# How recycling affects sustainable growth

Julie Metta<sup>a\*</sup> and L. Zhang<sup>b\*</sup>

<sup>a\*</sup>School of Energy and Environment, City University of Hong Kong, Hong Kong SAR
<sup>a</sup>Corresponding author: Julie Metta: j.metta@my.city.edu.hk - <u>https://orcid.org/0000-0001-</u>
<u>7141-8445</u> Tat Chee Ave, Kowloon Tong, City University of Hong Kong
<sup>b</sup>Lin Zhang - http://orcid.org/0000-0002-5945-5451 Email: l.zhang@cityu.edu.hk

#### Abstract

This paper quantifies the effect of waste recycling on GDP in OECD countries from 2000 to 2012. We focus on national levels and use external instruments to control for the endogeneity between GDP and waste recycling. Specifically, we examine the influence of labour, material flow and research and development (R&D) as drivers of this relationship. While R&D is often seen as key factor for a sustainable development, labour and material trades are of a higher importance for sustaining growth through recycling. Empirical results show a positive and statistically significant effect of waste recycling on the economy. An increase of 1% of the recycled waste raises GDP by up to 0.06%. Waste management, if well planned, can be a catalyst for a stagnating economy. Furthermore, our results show that the benefits of waste recycling are given rise through labour and resource flow channels. Governments are encouraged to initiate recycling investment accordingly, as such initiatives will not only improve the environment but also generate positive social welfare and pushes the transition to a circular economy.

Keywords: waste recycle; circular economy; sustainable development; resource.

#### 1 1. Introduction

2 Human activities generate waste, which increases mainly due to urbanization, population and industrialization (Andersen et al., 2007). The dominant part of waste generated, 3 Municipal Solid Waste (MSW)<sup>1</sup>, accounts for 5% of global greenhouse gas emissions (IPCC, 4 2014, 2007, 2006). In 2016, MSW was 2 billion tonnes and it is expected to rise to 2.2 billion 5 6 tonnes by 2025, according to the World Bank (Kaza et al., 2018). One possible method for 7 reversing this trend is through the implementation of a circular economy, where recycling is a 8 major component. As confirmed by Lacy and Rutqvist (2016) and Mavropoulos (2010), 9 efficient waste recycling could provide business opportunities and therefore increase Gross Domestic Product (GDP). In this paper, we study how waste<sup>2</sup> recycling affects economic 10 11 performance in developed economies.

12 Our study focuses on the OECD countries because developing countries have less mature 13 waste management policies and recycling systems, which makes them difficult to study. 14 Although waste recycling has drawn attention in many developed economies since the 1980s, 15 it is only recently that governments have imposed regulations for waste management. The European Union (EU) offers a circular economy package to sustain growth (European 16 Commission, 2010, 2015; FREC, 2018), and invested over €5.5 billion from 2004 to 2020 in 17 18 waste management in order to increase annual waste recycling capacity by 5.8 million tonnes 19 (Ten Brink et al., 2018). In Japan, a law was put into effect in 2001 to promote the efficient 20 utilization of resources. Its recycling industry contributes 7% to Japan's GDP and employs 21 0.65 million workers (METI, 2001). The economic effects of waste recycling strategies, 22 which are of importance to developed economies having implemented such policies, have

<sup>&</sup>lt;sup>1</sup> MSW accounts for all waste generated by the public. It agglomerates all domestic, streets and parks and gardens (green) waste. Business, office, institution and commercial waste similar – in term of kind and amount - to those produced by households are also assimilated to MSW.

<sup>&</sup>lt;sup>2</sup> Waste hereafter in the paper refers to municipal solid waste.

rarely been investigated empirically. This paper addresses this gap by evaluating theeconomic benefits of recycling initiatives.

25 The economic literature on waste recycling mainly focuses on the production and 26 management of the recycling industry. In the 1970s, studies mainly constructed economic models to understand the dynamics of waste accumulation and resource recycling (Smith, 27 28 1972; Schulze, 1974; Hoel, 1978; Slade, 1980), while recent studies have focused on the 29 related policy intervention for better waste management (Sidique et al., 2010; Jenkins et al., 30 2003; Shinkuma, 2003; Hamilton, 2003, Kinnaman and Fullerton, 2000). Another string of 31 the literature has tried to identify possible drivers of waste and recycling industries. 32 Economic development increases waste generation at the macro level (Ciacci et al., 2017; 33 Johnstone and Labonne, 2004), and household income has a positive and significant effect on 34 waste recycling at the micro level (Jenkins et al., 2003; Sidique et al., 2010). There is a recent 35 and growing literature on the role of household behaviours with regard to recycling (Alpízar and Gsottbauer, 2015; Abbott et al., 2017; Gilli et al., 2018). Limited research has 36 37 investigated the effects of waste recycling on the environment. Waste recycling can decrease greenhouse gas emissions, and thus strengthen atmospheric quality (Acuff and Kaffine, 38 39 2013). However, there is no strong evidence of the Environmental Kuznets Curve theory on 40 waste management (Mazzanti and Zoboli, 2009).

Although it is intuitive that waste recycling helps to save resources, empirical support for sustaining economic development by recycling has not been addressed in the literature. This paper fills this gap by quantifying the effect of waste recycling on the economic performance of developed economies. We use external instruments and apply panel econometric approach, based on data of 20 OECD countries over 13 years (2000-2012). Our results show that an increase of 1% of the recycled waste leads to an increase in GDP by up to 0.06% percent, implying that waste management, if well planned, can be a catalyst for a stagnating economy. 48 We further identify the underlying mechanisms that reinforce the benefits of waste recycling. While R&D is expected to be a driver, empirical results show that labour and material flow 49 50 are the two main channels. The recycling industry is identified as labour-intensive: higher 51 labour supply to the recycling sector enhances productivity (e.g. the quality of the recycled 52 material) and leads to higher economic output at country level (D'Amato et al., 2016; 53 Kinnaman et al., 2014). Our findings support the need for resource recovery policies and training programmes dedicated to workers in the recycling sector. Our results are robust to a 54 55 rich set of sensitivity checks addressing alternative waste and raw material indicators, reverse 56 causality, endogeneity and geopolitical considerations.

57 This paper contributes to the literature in a number of aspects. Firstly, the existing 58 literature linking waste management and growth focuses on the microeconomic aspects of 59 recycling (Mazzanti and Zoboli, 2008, 2009). Information on the macroeconomic effects of 60 waste recycling practices is generally missing. This paper is the first to empirically estimate 61 the impact of waste recycling on economic outputs for developed countries. In addition, we 62 explore new mechanism through which the positive effects of waste recycling on economic output can be enforced. The new perspective highlights the need for engaging labour force 63 64 into the recycling industry rather than deploying investment in R&D.

Secondly, our paper complements the literature on economic growth. Traditional growth 65 66 models (Romer, 1990; Aghion and Howitt, 1992; Helpman, 1992; Barro, 1996; Bretschger et 67 al., 2017) have highlighted that population growth, R&D and knowledge diffusion are the 68 driving forces for sustainable growth. Di Vita (2001), Pittel et al. (2010), and Fagnart and 69 Germain (2011) confirm, from the theoretical perspective, recycling as another driver for growth. On the one hand, our results support this argument. On the other hand, our results 70 71 show no evidence of R&D as underlying mechanism but highlight labour and material as 72 main drivers to sustain development through recycling.

73 Thirdly, this study addresses the co-benefit analysis of environmental policies. 74 Environmental economists are concerned about the degradation of natural resources and 75 ecosystems and about the impacts of climate change. However, the policies to mitigate the 76 environmental impacts are highly costly and often represent a significant proportion of 77 government budgets (Almer and Winkler, 2017; Burtraw et al., 2014). Managing waste 78 through recycling should be an effective way to reduce the cost of environmental policies, as 79 effective waste recycling lowers greenhouse gas emissions and prevents environmental 80 degradation due to excess resource extraction and consumption (Hamilton et al., 2013; Acuff 81 and Kaffine, 2013). Recycling can therefore be considered a global and competitive strategy 82 for climate change mitigation.

The remainder of the paper is organized as follows. Section 2 introduces our empirical model and explains our econometric strategy. It also details the choice of our instruments and their validity. Section 3 describes the data and provides descriptive statistics, and Section 4 presents our main results. Section 5 provides a series of robustness checks by using alternative measures and different estimators. Section 6 identifies different mechanisms driving the relationship between recycling and economic outcomes. Section 7 concludes the paper and highlights potential policy implications.

90

# 2. Empirical model and econometric method

91 **2.1. The empirical model** 

The model is developed following the theoretical literature on growth and recycling in developed countries. We consider a standard empirical growth model of economy-wide production with technological change (DiVita, 2001; Pittel et al., 2010, Fagnart and Germain, 2011). In our model, the aggregate output *Y* is produced by using labour (*L*), physical capital (*K*), human capital (*H*), resource flows (or material input) (*M*), the technology level (*A*) and the amount of recycled waste (*W*) as production inputs. Consistent with the empirical 98 literature on cross-country growth models (Kormendi and Meguire, 1985; Acemoglu et al.,
99 2001), with all of the variables being transformed into a log-linear specification, the model
100 takes a simple form reflecting the relationships as follows:

$$\ln(Y_{i,t}) = \beta_W \ln(W_{i,t}) + \beta_i \boldsymbol{\mathcal{X}}_{i,t} + \varepsilon_{i,t} , \#(1)$$

101 where  $W_{i,t}$  is the measure of waste recycling,  $\beta_W$  is the coefficient of interest,  $\boldsymbol{X}_{i,t}$  is a vector 102 of other covariates<sup>3</sup>,  $\beta_i$  is the unknown parameters to be estimated, and  $\varepsilon_{i,t}$  is the unobserved 103 error term. We expect waste recycling to have a positive impact on the economic outputs. By 104 including all the controls, the model to be estimated reads as follows:

$$ln(Y_{i,t}) = \beta_W ln(W_{i,t}) + \beta_1 ln(K_{i,t}) + \beta_2 ln(L_{i,t}) + \beta_6 ln(H_{i,t}) + \beta_5 ln(M_{i,t}) + \beta_7 ln(A_{i,t}) + \epsilon_{2i,t}. \#(2)$$

105 Following Verdolini and Galeotti (2011), government investment in research and 106 development (R&D) is used as proxy for the provincial specific technological changes (A), 107 which could capture technology heterogeneity. The existing theoretical literature on the effect 108 of waste recycling on growth suggests a negative effect of technological change on recycling 109 (Di Vita, 2001). In this sense and to consider a neutral technical change common to all countries, the function for A is additionally specified as a linear function of time t and  $t^2$ . To 110 further investigate the impact of such a trend on the effect of waste recycling on economic 111 112 performance, the model is revised as follows:

113 
$$ln(Y_{i,t}) = \beta_W ln(W_{i,t}) + \beta_1 ln(K_{i,t}) + \beta_2 ln(L_{i,t}) + \beta_6 ln(H_{i,t}) + \beta_5 ln(M_{i,t}) + \beta_7 ln(A_{i,t}) + \beta_3 t + \beta_4 t^2 + \varepsilon_{3i,t}. \#(3)$$

We first apply the fixed effects (FE) estimator to estimate equations (2) and (3).<sup>4</sup> All model specifications include country-fixed effects to capture the effects of within-country changes. The transmission channels are identified and further investigated in section 6.

<sup>&</sup>lt;sup>3</sup> The covariates are assumed to be exogenous – see results Table A 1 in the appendix A.

<sup>&</sup>lt;sup>4</sup> The Hausman test shows that the fixed effects model is preferred to the random effects model, which confirms the homogeneity of the OECD data detailed in Section 3.

#### 117 **2.2. Instrumental variables**

Our model may suffer from endogeneity issue due to potential reverse causality. To precisely estimate the impact of waste recycling on the economy, we need to address the endogeneity issues between GDP and waste recycled. We use a two-stage least squares method (2SLS) with external instrumental variables (IVs) as proxy for recycled waste. Therefore, it is necessary to estimate the following equations, where equations (5) and (6) represent the first and second stages of this approach, respectively:

$$\ln(W_{i,t}) = \alpha_{IV} \ln(IV_{i,t}) + \alpha_i \mathbf{X}_{i,t} + \nu_{i,t} \#(4)$$
$$\ln(W_{i,t}) = \alpha_{2_{IV}} \ln(IV_{i,t}) + \nu_{2_{i,t}} \#(5)$$
$$\ln(Y_{i,t}) = \beta_{IV} \ln(IV_{i,t}) + \beta_{2_i} \mathbf{X}_{i,t} + \varepsilon_{IV_{i,t}}, \#(6)$$

where *IV* represents the measure used as instrument variable (IV),  $\alpha_{IV}$  and  $\alpha_{2_{IV}}$  are the coefficients of the *IV* to be estimated, and  $\nu$  and  $\nu_2$  represent the error terms.

Our estimation uses two external IVs for recycled waste (W): environmental policy stringency (EPS) and the number of people connected to a water sewage grid (Water). This identification requires that IV is not correlated with the error term. In what follows, we show that this assumption is validated for our instrument choices.

The environmental policy stringency variable refers to the Environmental Policy Stringency Index, which is a country-specific and internationally comparable measure of the stringency of the environmental policy. The index ranges from 0 (not stringent) to 6 (highest degree of stringency). The level of stringency is based on the degree of stringency of 14 environmental policy instruments.



Figure 1 - Environmental Stringency Index of selected countries between 2000 and 2012
Source: Botta and Kozluk (2014)

135

138 The rationale behind the correlation between EPS and waste can be explained as follows. 139 Waste recycling is considered as one of the best choices for governments to improve their 140 environmental policy stringency. This is because regulations on waste management are easier 141 to implement and less costly to set than other environmental policies such as emission reduction (Hamilton et al., 2013; Acuff and Kaffine, 2013). Recycling policies are also rapid 142 143 to set and implement. On the other hand, recycling-related environmental policy has direct 144 and immediate effect on quantity of recycled waste, while other policies that regulate human 145 impacts on the environment may take more than one year to be implemented and even longer 146 to demonstrate a significant impact on economic performance. It is likely to take years from such policy design to increased environmental technology research, before the final 147 148 utilization of such innovation on managing waste. Therefore, EPS in general is less correlated 149 with GDP over the period in this study. As confirmed by Albrizio et al. (2017), there is not 150 any strong relation between economic growth and EPS at the national level, which supports 151 our rationale for using EPS as an instrument for waste recycling.

152 The second instrument is waste water treatment (Water). The indicator "waste water treatment"<sup>5</sup> represents the percentage of the population connected to a waste water treatment 153 plant through a public sewage network. It does not take into account independent private 154 155 facilities that are used when public systems are not economical. Studies have shown that waste and water are two resources similarly managed and very much interconnected 156 (Hamilton et al., 2013; Kellenberg, 2012; Harford, 1976). As "water supply" is often used as 157 an instrument for waste, similarly, waste water treated could be used as an instrument for 158 159 waste recycled. Moreover, no study provides a direct nexus between water effluent 160 management specification and economic growth. In OECD countries, sludge is historically 161 managed by regional public institutions. Also, many towns have no networks and households 162 own independent private sludge systems. In France, more than 5 million households are not 163 connected to any urban network. In the European Union, networks have only been 164 compulsory since 1991 and for population centres with over 2000 inhabitants. The time 165 period studied comes after this regulation was introduced. Figure 2 shows that even though 166 Ireland, Portugal and the Czech Republic are part of the European Union, they have experienced different variations in the period studied. Waste water treatment data are 167 168 uncorrelated with the error term.

<sup>&</sup>lt;sup>5</sup> Water hereafter in the paper refers to waste water treatment.



169

170 Figure 2 - Waste water treatment of selected countries between 2000 and 2012 Source:

171

#### OECD (2018e)

172 In addition, we perform a series of statistical tests where EPS, Water or both are used as 173 instruments to ascertain the extent to which our IVs are valid in controlling for the 174 endogeneity. In the first stage regressions, the Kleibergen-Paap Wald F-tests are conducted 175 to see if the IVs are jointly significant for the endogenous variable, W. All the results of the 176 first stage of the 2SLS are summarized in Table A 3 in the appendix A. In the first stage regressions, all F statistics are bigger than 10, confirming the strength of our IVs to 177 178 instrument W. The Kleibergen–Paap Lagrange multiplier test (LM) rejects the null hypothesis of under-identification at 1% significance<sup>6</sup>. The Cragg-Donald Wald F statistics are greater 179 than the Stock and Yogo (2005) at the 10% level. Comparison of the Kleibergen–Paap Wald 180 181 test with the critical values for the Cragg-Donald statistic in the second stage (Bazzi and Clemens, 2013) validates the strength of the chosen IVs. The Hansen J statistic, which is used 182

<sup>&</sup>lt;sup>6</sup> This diagnosis assesses the identification of the equation and allows us to reject the null hypothesis of the equation being under-identified. The 2SLS with IV approach is appropriate and valid for all models.

183 to check for overidentification of the instruments, shows that IVs are significantly correlated

184 with *W*, but not with disturbances.

185 The results from the above diagnostics support the use of these instruments jointly to

186 control for the endogeneity of *W* in the regressions.

187

Table 1 – Statistical test results of the 2SLS approach

	Water	&EPS	E	PS	Water		
	No time		No time		No time		
	trend	Time trend	trend	Time trend	trend	Time trend	
	(1)	(2)	(3)	(4)	(5)	(6)	
lnW	0.0536***	0.0425***	0.2524*	0.1845	0.0327	0.0422**	
	(0.0203)	(0.0185)	(0.1678)	(0.3975)	(0.0205)	(0.0185)	
Endogeneity test	0.224	0.53	8.715	0.431	0.297	0.075	
$Prob > chi^2$	0.6357	0.8183	0.0032	0.5116	0.5857	0.7836	
KP LM stat. (under id)	18.092***	18.549***	1.657	0.218	17.754***	17.999***	
KP Wald F stat. (weak id)	13.126	15.288	1.71	0.211	24.573	29.272	
	19.93(10%)	19.93(10%)	16.38(10%)	16.38(10%)	16.38(10%)	16.38(10%)	
CD Wald F stat. (weak id)	11.59(15%)	11.59(15%)	5.53(25%)	5.53(25%)	5.53(25%)	5.53(25%)	
Hansen J statistic (over id)	8.651	0.406	-	-	-	-	
$Prob > chi^2$	0.033	0.5241	-	-	-	-	

188Note: The results are estimated with 2SLS. The dependant variable is  $\ln GDP$ . Country and year fixed effects are189included. All test statistics in the table are robust to heteroscedasticity and arbitrary within-correlation. KP holds for190the Kleibergen and Paap (2006) tests. They account for underestimation and weak identification instruments (Bazzi191and Clemens, 2013). CD refers to the Cragg-Donald test. The columns with odd index are for regressions that do not192include any "time trend" while those with even index for regressions do include a time trend. \* p<0.10, \*\*p<0.05,</td>193\*\*\*p<0.01. Standard errors are reported in brackets.</td>

Furthermore, the Wooldridge and regression-based tests are performed to check the endogeneity of *W*. The Wooldridge and regression-based tests allow for heteroskedastic and autocorrelated errors. The endogeneity test in Table 1 rejects the hypothesis that *W* is 197 exogenous. All the results from the above diagnostics confirm the relevance and strength of198 the external IVs.

#### 199 **3. Data**

The data used in this analysis are combined from various sources to construct a unique balanced country-level panel dataset on the development of 20 OECD<sup>7</sup> countries from 2000 to 2012. The OECD countries are chosen for their similarity in environmental policies. As depicted by Figure 3, the selection offers different geographical characteristics in the developed world.



205 206

Figure 3 - Countries included in this study



<sup>&</sup>lt;sup>7</sup> 20 countries were chosen from among all OECD countries. If the availability of data has driven the selection of the countries for control and instruments variables, these countries sufficiently represent the OECD countries, given that even if they are only 59% of the countries, they jointly account for 70% of its population and 77% of its GDP in 2012.

Monitory Fund (IMF 2017) database<sup>8</sup> (Kamps, 2006; Gupta et al., 2014). Human Capital data were retrieved from World Bank (2018). The data for the IV EPS was collected from Botta and Kozluk (2014). The data sources are further described in Table B 1 in the Appendix B. Figure *4* - Statistics on variables used as dependent and control variablesFigure 4 shows the descriptive statistics on the variables used in the study.



216

Figure 4 - Statistics on variables used as dependent and control variables Sources: OECD
(2018a,b,c,d,e), UN (2017), UNEP (2016), IMF (2017), Kamps, (2006) Gupta et al. (2014),
World Bank (2018).

GDP is used to represent a country's economic output Y, which is measured in million constant USD at 2010 prices and corrected for PPP exchange rates. The explanatory variable of interest is waste recycled (W), in total kilograms per country. Waste recycled is the measured weight of municipal waste recycled. Figure 5 a illustrates that for most of the

<sup>&</sup>lt;sup>8</sup> This variable is calculated using the IMF Investment and Capital Stock Dataset, 2017, which extended the methodology of Kamps (2006) and Gupta et al. (2014). The IMF (2017) file provides comprehensive data on public investment and capital stock (i.e. general government), private investment and capital stock, as well as investment and capital stock arising from public-private partnerships (PPPs), across the Fund membership countries.

countries in Europe waste recycled amount increase while in Asia the aggregate amount decreases over the studied period. Figure 5 b highlights the large variation of recycling waste over time and countries. Starting from a very low level of recycling, Poland had a rise of over 809% from 2000 to 2012, while the average increase in the countries studied was 132%.





229

Figure 5 - Waste recycled in selected countries Source: OECD (2018,b)

Figure 6 shows a clear and strong positive relationship between GDP and waste recycled. Countries with high economic development are generally places where the recycling activities are intensive. Our theory is that this relationship reflects the effect of waste recycling through an improved sustainable growth mechanism.



Figure 6 - Scatter plot of per capita GDP and waste recycled Source: OECD (2018a,b)

234

Note: The figure plots the logarithm of GDP per capita against the logarithm of the municipal solid
waste being recycled for the sample analysed. The linear fitted trend is arbitrary. The purpose of this
paper is to analyse whether such a trend exists, linking the two variables.

To evaluate the relevance of our hypothesis and to limit the chance of omitted variables, we include different control variables in our regressions. The capital stock (K) is expressed by the sum of public and private capital. Labour (L) is the labour resource of workers, who are people available for work among the working age population.

To account for resource flows (M), we use Direct Material Input (DMI)<sup>9</sup>. DMI measures the total amount of material directly used within an economy. It includes biomass, metal, minerals, and fossil fuels and is defined as the weight of domestically extracted raw materials, plus direct material imports and direct material exports.

Human Capita (*H*) is captured by government expenditures (in real terms) dedicated to tertiary students. Government investment in research and development (R&D) is used as a proxy for the technological variable.

<sup>&</sup>lt;sup>9</sup> A robustness check is undertaken in Section 5 using DMC as alternative data for resource flows.

To control for concerns on macroeconomic series characteristics, all data were checked for unit root (Nelson, 1982; Stock and Watson, 1989). Following Dickey and Fuller (1979), augmented Dickey-Fuller tests (ADF) are applied across all control variables to confirm the stationarity of the cross-sectional panel data. The results of the ADF unit root tests for levels and first differences show that in all countries, the variables are stationary. The results are detailed in Appendix A.

#### **4. Results**

Table 2 presents the estimated results for our model. The specifications of equations (2) and (3) are entitled "time trend" and "no time trend", respectively. For each of the specifications, both ordinary least squared fixed effects model (FE) and Two-Stage Least Gquares regressions with instruments (2SLS) are presented.

Overall, we observe that there is a strong positive and statistically significant effect of waste recycling on economic output. The results obtained with FE and 2SLS can be considered similar. For the fixed effects model, the results for both specifications are 0.044 and 0.037 (columns (1) and (3), respectively), while with 2SLS, the coefficients are 0.054 and 0.043 in columns (2) and (4), respectively. By comparing the results with and without time trend, we find that the estimated coefficients are smaller when time trends are included.

The magnitude of the coefficient of waste is larger in the 2SLS specification compared to the FE. This is probably due to an inherent measurement error bias in FE estimators that underestimates the effect of waste recycling on economic outputs. The 2SLS estimator is considered to provide the more solid results. These findings suggest that an increase in 1% in waste recycled will lead to an increase of up to 0.05% in GDP.

In all regressions, the estimated coefficients for all the control variables are consistent and statistically significant. Material (resource flows) and labour have high (above average) positive effects on economic output as they are the two major inputs for economy-wide production. The coefficients show that a 1% increase in labour contributes to a more than 0.3% increase in GDP, while the effect of material is over 0.2%. Both physical and human capital are positive and statistically significant for GDP. The effect of physical capital is much higher (by 39%) than human capital, with estimated coefficients up to 0.111 and 0.067, respectively.

Table	2 –	Estimation	results

	No tin	ne trend	Time trend		
	FE	2SLS	FE	2SLS	
	(1)	(2)	(3)	(4)	
lnW	0.0443***	0.0536***	0.0372***	0.0425***	
	(0.0082)	(0.0203)	(0.0079)	(0.0185)	
ln <i>K</i>	0.115***	0.111***	0.0776***	0.0759***	
	(0.0206)	(0.0249)	(0.0213)	(0.0210)	
ln <i>L</i>	0.358***	0.348***	0.335***	0.330***	
	(0.0628)	(0.0790)	(0.0599)	(0.0804)	
ln <i>M</i>	0.242***	0.247***	0.244***	0.247***	
	(0.0187)	(0.0205)	(0.0192)	(0.0215)	
ln <i>H</i>	0.0707***	0.0670***	0.0468***	0.0453***	
	(0.0143)	(0.0162)	(0.0144)	(0.0150)	
lnA	0.117***	0.115***	0.112***	0.111***	
	(0.0194)	(0.0187)	(0.0193)	(0.0201)	
t	-	-	0.0117***	0.0114***	
	-	-	(0.0023)	(0.0026)	
$t^2$	-	-	-0.000551***	-0.000541***	
	-	-	(0.0001)	(0.0002)	
Country FE	yes	yes	yes	yes	
Adj R <sup>2</sup>	0.89	0.889	0.902	0.901	
Ν	260	260	260	260	

F	354.7	285.8	300	251.1
---	-------	-------	-----	-------

281	Note: The results are estimated with FE and 2SLS. The dependent variable is lnGDP. The Hausman test
282	was conducted to confirm the validity of the choice for FE. For the 2SLS, EPS and Water are used jointly as
283	external instrument variables. All test statistics in the table are robust to heteroscedasticity and arbitrary within-
284	correlation. * p<0.10, **p<0.05, ***p<0.01. Values in parenthesis are the standard errors.

The estimated coefficient for R&D is higher than for those of physical capital, implying that technological innovation is one of the main drivers of growth in developed countries. The time trend evidences the quadratic impact of technological changes on economy.

The positive sign of the estimates of t and negative sign of those of  $t^2$  explain that the general technical improvement is increasing at a decreasing rate. Therefore, the impact of general technology improvement, which is captured by the time trend, shows a strong positive effect on economic performance.

292 The effect of waste recycling on GDP can be argued to be low compared to the other 293 variables. According to Pittel et al. (2010), this low level of impact can be explained by the 294 lack of investments in the recycling industry. Among OECD countries, on average, only 295 1.46% of material used in the economy (according to the DMI data) is reinjected through the 296 recycling of municipal solid waste. This ratio is particularly low in Turkey, representing only 297 0.001% of the material initially injected. Switzerland has the highest rate of recycling of the 298 developed countries studied in this paper, with a rate of 5.5%. The effects of recycling waste 299 on GDP are positive for all countries.

Figure 7(a) illustrates the marginal effects of waste recycling on GDP across countries. It confirms the effects of recycling waste on GDP are positive and similar for all countries. This can be explained by the commonly ratified international agreements and standardised level of environmental awareness among countries. Figure 7(b)) shows the deviation of country individual effect from the mean value of all countries. We observe that the deviation is positive for the first few years and then becomes negative for European countries, but not for 306 non-European countries like Japan. One explanation for this is the change of leadership. In 307 the early 21<sup>st</sup> century countries in Europe were the leaders in waste management (Milligan 308 and O'Keeffe, 2019). As non-European countries start to improve their regulations on waste 309 trade and management, they take the leadership from European countries in this field.



310

311 Figure 7 - Marginal effect of waste recycling on GDP by country

Note: Panel (a) shows the marginal effect of waste recycling on GDP for 6 selected countries and the averagevalues. Panel (b) shows the deviation of country individual effect from the mean value of all countries.

Our analysis shows evidence of the positive effect of waste recycling on economic performance at the national level and confirms empirically the theoretical work on endogenous growth undertaken by Di Vita (2001). These results also complement to the literature that highlights the positive impact of waste recycling at within-country levels (Mazzanti and Zoboli, 2008; Slade, 1980) and at microeconomic level for Italy (Mazzanti and Zoboli, 2008).

#### **5. Robustness Check**

To assert the validity of our results, we conduct a series of robustness checks. Our findings are robust to a rich set of sensitivity checks addressing alternative waste and material indicators, reverse causality, endogeneity and geopolitical considerations. We now explainand show each of the analyses.

325

#### 5.1. Effects of recycling rate, long term policies and the role of households

Three alternative measures of waste recycling are considered. The first model variation introduces one period lag. This change also controls for endogeneity and estimates selfaccumulation of the variables. Our second alternative test is to turn each variable into per capita value. This variation allows us to estimate the role of households in the recycling sector. Our third alternative is to use the share of waste recycled over the total amount of waste collected as a variable to account for *W*. This proxy can be interpreted as the recycling rate.

333 Table 3 displays the results of three alternative specifications with both the FE (columns 334 noted as a) and the 2SLS approach (columns noted as b). All columns with odd numbers (1, 3, 5) do not consider a time trend, while columns with even index (2,4, 6) consider a time 335 336 trend. These results are all significant and consistent with the original model (results Table 2), 337 which confirms the validity of our results. Columns 1 and 2 of Table 3 show the model 338 results when one period lag of W is introduced. We find that the effect of recycled waste on 339 economic performance is higher (at 12.7%) when estimated with the 2SLS approach 340 compared to our benchmark results. As we use environmental policy stringency (EPS) as an 341 instrument for recycled waste (W), this first variable is taken with one period lagged. We can 342 consider that the repercussions of recycling regulations are even stronger one year after being 343 implemented. Thus, once a recycling policy is enforced, the effects are enhanced the year 344 later, leading to a stronger effect of recycling in this alternative approach. This finding 345 validates the theory first developed by Jenkins et al. (2003) on the need for long-term 346 recycling policies.

The third and fourth columns of Table 3 summarizes the results when data are accounted for by per capita values. The effect of recycled waste on GDP per capita is up to 0.0624, namely, 1% increase of per capita waste recycled will result in a 0.062% increase in GDP. This suggests that the positive effect of waste recycling is not affected by the population size. Therefore, governments should legislate the enhancement of waste recycled per person or at household level, as it further raises GDP (16.4% more than with aggregate recycling).

The last columns (5, 6) of Table 3 exhibits the results when the share of waste recycled over the total amount of waste collected is used. It shows that a one unit increase of waste recycling rate contributes up to a 5.66% increase in GDP. Increasing the waste recycling ratio increases economic outputs at a higher rate (about 6.5% higher) compared to increasing the total amount of waste recycled. This result provides empirical support for waste reduction policies and encourages higher market penetration in the waste recycling industry.

359 Our main findings hold and strengthen the need for policies both for waste reduction and 360 increasing the recycling rate, when using different measures of waste recycled.

	Lag W				Per capita data			Share of waste recycled				
	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS
	(1a)	(1b)	(2a)	(2b)	(3a)	(3b)	(4a)	(4b)	(5a)	(5b)	(6a)	(6b)
$\ln W$	0.0359***	0.0604***	0.0299***	0.0533***	0.0473***	0.0624**	0.0416***	0.0461*	0.0455***	0.0566**	0.0400***	0.0478**
	(0.0092)	(0.0213)	(0.0085)	(0.0171)	(0.0084)	(0.0264)	(0.0080)	(0.0244)	(0.0085)	(0.0244)	(0.0081)	(0.0216)
ln <i>K</i>	0.132***	0.113***	0.0766***	0.0620**	0.107***	0.0978***	0.0644***	0.0624***	0.119***	0.115***	0.0790***	0.0770***
	(0.0256)	(0.0336)	(0.0267)	(0.0310)	(0.0205)	(0.0253)	(0.0217)	(0.0223)	(0.0205)	(0.0251)	(0.0211)	(0.0210)
lnL	0.338***	0.319***	0.318***	0.301***	0.299***	0.265***	0.304***	0.294***	0.397***	0.394***	0.366***	0.365***
	(0.0661)	(0.0941)	(0.0612)	(0.0954)	(0.0673)	(0.0998)	(0.0640)	(0.0949)	(0.0623)	(0.0769)	(0.0591)	(0.0787)
$\ln M$	0.242***	0.258***	0.246***	0.261***	0.248***	0.259***	0.249***	0.252***	0.236***	0.241***	0.240***	0.243***
	(0.0194)	(0.0225)	(0.0197)	(0.0222)	(0.0191)	(0.0230)	(0.0194)	(0.0243)	(0.0185)	(0.0203)	(0.0189)	(0.0209)
ln <i>H</i>	0.0622***	0.0553***	0.0395***	0.0342**	0.0705***	0.0650***	0.0458***	0.0445***	0.0633***	0.0572***	0.0390***	0.0353**
	(0.0154)	(0.0168)	(0.0149)	(0.0157)	(0.0140)	(0.0167)	(0.0144)	(0.0155)	(0.0147)	(0.0195)	(0.0147)	(0.0171)
lnA	0.120***	0.112***	0.116***	0.109***	0.114***	0.109***	0.111***	0.110***	0.122***	0.121***	0.117***	0.116***
	(0.0207)	(0.0188)	(0.0197)	(0.0198)	(0.0194)	(0.0198)	(0.0192)	(0.0207)	(0.0193)	(0.0181)	(0.0191)	(0.0190)
Time trend	No	No	Yes	Yes	No	No	Yes	Yes	No	No	Yes	Yes
N. of obs	240	240	240	240	260	260	260	260	260	260	260	260
$\mathbb{R}^2$	0.878	0.874	0.896	0.892	0.839	0.836	0.855	0.855	0.89	0.889	0.902	0.902
F	292	220.1	261.4	220.4	229.4	198.9	194.7	171.6	353.6	280.7	302.6	247.4
KP LM (under id) KP Wald F		15.974		15.549		13.306		12.434		15.676		17.179
(weak id)		10.825		11.191		9.611		10.507		11.305		13.595
Hansen J		11.717		4.676		8.096		0.134		9.739		0.746
$Prob > chi^2$		0.0006		0.0306		0.0044		0.7143		0.0018		0.3878
Endog. test		1.05		1.369		0.273		0.021		0.164		0.093
Prob > chi <sup>2</sup>		0.3054		0.242		0.6012		0.8861		0.6858		0.7598

361 Table 3 - Results with alternative measures of recycled waste (*W*)

362 363 364 Note: The table reports estimates for the regressions using different data to measure W only (one period lag of W in columns 1 & 2 and share of MSW recycled over the total of MSW treated for 5 & 6) and per capita data for all the variables (columns 3 & 4). The dependent variable is lnGDP. The columns with odd index are for regressions that do not include any "time trend", while those with even index are for regressions that do include a time trend. The Hausman test was conducted to confirm the validity of the choice for FE. For the 2SLS, EPS and Water are used 365 jointly as external IVs. For all regressions estimated with 2SLS, the Cragg-Donald Wald F stat. (weak id) is 19.93 at 10% and 11.59 at 15%. All test statistics in the table are robust to 366 within-correlation. \* p<0.10, \*\*p<0.05, \*\*\*p<0.01. Values in heteroscedasticity and arbitrary parenthesis are the standard errors

#### 367 **5.2.** Effects of raw material imports

The alternative indicator used to analyse the resource flows is DMC. This indicator is similar to DMI but does not include importations of materials (DMC = DMI – imports). All regressions yielded results – summarized in Table 4 - nearly identical to those for DMI in our benchmark model (results Table 2). All estimates are consistent and significant at least at the 10% level.

2	7	2
Э	1	Э

Table 4 - Results with alternative measures of material (*M*)

		Using DMC	for Material	
	FE	2SLS	FE	2SLS
	(1a)	(1b)	(2a)	(2b)
$\ln W$	0.0434***	0.0485**	0.0363***	0.0395*
	(0.0087)	(0.0220)	(0.0084)	(0.0210)
$\ln M$	0.199***	0.202***	0.200***	0.202***
	(0.0181)	(0.0194)	(0.0186)	(0.0218)
ln <i>K</i>	0.120***	0.117***	0.0819***	0.0809***
	(0.0219)	(0.0257)	(0.0227)	(0.0214)
ln <i>L</i>	0.320***	0.313***	0.294***	0.290***
	(0.0696)	(0.0848)	(0.0666)	(0.0871)
ln <i>H</i>	0.0775***	0.0757***	0.0541***	0.0533***
	(0.0154)	(0.0162)	(0.0154)	(0.0150)
lnA	0.123***	0.122***	0.121***	0.120***
	(0.0206)	(0.0204)	(0.0204)	(0.0222)
Time trend	No	No	Yes	Yes
N. of obs	260	260	260	260
$\mathbb{R}^2$	0.876	0.875	0.889	0.888
F	309.3	253.8	261.7	206.3
KP LM		17.184		17.709
KP Wald F		12.186		13.984
Hansen J		6.951		0.075
$Prob > chi^2$		0.0084		0.7844
Endog. test		0.038		0.018
$Prob > chi^2$		0.8458		0.8947

	0.0450 0.0747
374	Note: The table reports estimates for the regressions using different data to measure M only (DMC). For
375	all regressions estimated with 2SLS, the Cragg–Donald Wald F stat. (weak id) is of 19.93 at 10% and of
376	11.59 at 15%. All test statistics in the table are robust to heteroscedasticity and arbitrary within-
377	correlation. * p<0.10, **p<0.05, ***p<0.01. Values in parenthesis are the standard errors

The effect of waste recycling on GDP is lowered by 5% (2% for FE and 8% for 2SLS) when DMC is used instead of DMI as proxy for resources. Material imports also have a 5%

effect on the GDP. <sup>10</sup> The results highlight the role of the imports of raw materials as a
transmission channel for waste recycling, as detailed in Section 6.

382

#### 5.3. Geopolitical considerations

To address the potential biases attributed to geopolitical diversity, we construct different sub-samples of the dataset. We cluster the countries alternatively into three different geopolitical groups to evaluate if our estimated elasticities are valid. The results are reported in Table 5. In most cases, the estimated elasticities of waste are consistent with the benchmark model and significant.

388 Firstly, we cluster the countries members of the European Union (EU), as EU 389 members are required to follow similar regulations that may affect their economic behaviours 390 and outputs. The first columns (1 and 2) suggest that among EU members, a 1% increase in 391 MSW recycling increases GDP by up 0.088%. The second group is made up of countries 392 under the Schengen agreement and the third group consists of countries in the Euro zone. The 393 results indicate that the positive effect of waste recycling on GDP is higher compared to the 394 benchmark results. A possible reason for this result could be that the opportunities for free travel within the Schengen region helps to spread waste recycling technology and social 395 396 behaviours. Also, the Euro zone countries lead in the recycling markets and technological 397 advancement. These countries probably merge and encourage others to improve their 398 recycling systems. Therefore, the positive results are possibly enabled by the core structure of 399 the EU, Schengen area and Euro zone. It results that in Europe, the implementation of 400 policies on waste recycling would have stronger effects on the national economic 401 performance than other OECD countries. in

<sup>&</sup>lt;sup>10</sup> The calculation was done by comparing the mean of the difference between DMI and DMC over DMI and multiply it by the effect of M on GDP. The effect of recycled waste (W) over GDP is lessened by 5.3% while the importations are responsible for 4.8%.

		Using the sa	ample for EU		U	sing the sample	for Schengen A	rea		Using the sample for Euro zone		
	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS
	(1a)	(1b)	(2a)	(2b)	(3a)	(3b)	(4a)	(4b)	(5a)	(5b)	(6a)	(6b)
$\ln W$	0.0371***	0.0593**	0.0423***	0.0883***	0.0431***	0.0507*	0.0496***	0.0833***	0.011	0.0714**	0.0137	0.0693***
	(0.0109)	(0.0236)	(0.0109)	(0.0234)	(0.0106)	(0.0261)	(0.0105)	(0.0239)	(0.0144)	(0.0284)	(0.0143)	(0.0244)
ln <i>K</i>	0.165***	0.135***	0.117***	0.0385	0.139***	0.129***	0.0918**	0.0441	0.150***	0.0403	0.162***	0.0685
	(0.0309)	(0.0425)	(0.0407)	(0.0483)	(0.0286)	(0.0401)	(0.0355)	(0.0401)	(0.0423)	(0.0671)	(0.0503)	(0.0520)
ln <i>L</i>	0.441***	0.457***	0.450***	0.486***	0.455***	0.464***	0.477***	0.521***	0.532***	0.544***	0.501***	0.503***
	(0.0741)	(0.0913)	(0.0735)	(0.0946)	(0.0784)	(0.0964)	(0.0767)	(0.1030)	(0.0802)	(0.0998)	(0.0853)	(0.0926)
ln <i>M</i>	0.252***	0.245***	0.231***	0.213***	0.280***	0.277***	0.250***	0.235***	0.195***	0.216***	0.187***	0.206***
	(0.0236)	(0.0258)	(0.0247)	(0.0250)	(0.0240)	(0.0273)	(0.0249)	(0.0271)	(0.0290)	(0.0280)	(0.0315)	(0.0282)
ln <i>H</i>	0.0468***	0.0382**	0.0386**	0.0186	0.0402***	0.0367*	0.0264*	0.00935	0.0507***	0.0528***	0.0548**	0.0590***
	(0.0157)	(0.0185)	(0.0160)	(0.0188)	(0.0152)	(0.0215)	(0.0153)	(0.0205)	(0.0193)	(0.0198)	(0.0217)	(0.0206)
lnA	0.112***	0.105***	0.113***	0.0975***	0.116***	0.114***	0.116***	0.105***	0.0970***	0.0952***	0.104***	0.104***
	(0.0203)	(0.0192)	(0.0207)	(0.0207)	(0.0190)	(0.0188)	(0.0195)	(0.0202)	(0.0207)	(0.0152)	(0.0213)	(0.0178)
Time trend	No	No	Yes	Yes	No	No	Yes	Yes	No	No	Yes	Yes
N. of obs	195	195	195	195	208	208	208	208	130	130	130	130
$\mathbf{R}^2$	0.893	0.89	0.896	0.885	0.89	0.89	0.897	0.891	0.88	0.86	0.881	0.864
F	273	247.4	212.3	202.3	283.9	236.9	228.2	192.5	159.5	162.2	121.9	131.9
KP LM		8.464		7.124		7.037		5.963		9.361		12.984
KP Wald F		13.851		9.79		8.997		7.231		13.631		19.218
Hansen J		2.03		0.01		1.391		0.005		0.313		0.051
Prob > chi <sup>2</sup>		0.1542		0.919		0.2383		0.9464		0.5759		0.822
Endog. test		1.021		3.404		0.161		1.609		6.22		9.466
Prob > chi <sup>2</sup>		0.3123		0.065		0.6887		0.2047		0.0126		0.0021

402Table 5 - Results with countries clustered

403 Note: The table reports estimates for the regressions clustering different countries (European Union 1 & 2, Schengen area for columns 3 & 4 and lastly Euro zone for

404 columns 5 & 6). All test statistics in the table are robust to heteroscedasticity and arbitrary within-correlation. \* p<0.10, \*\*p<0.05, \*\*\*p<0.01. Values in parenthesis are the

405 standard errors

#### 406 **5.4.** Alternative control for endogeneity

407 We use the Hausman and Taylor (1981) (HT) with the Amemiya-MaCurdy (AM) 408 specification to control for endogeneity. By modifying the HT estimator, Amemiya and 409 MaCurdy (1986) developed an instrumental variable estimator to control for individual-410 specific unobservable effects that may be correlated with other explanatory variables. 411 Additionally, the HT AM method is robust under heteroskedasticity. To verify the 412 consistency of the HT AM estimator and thus the validity of the instruments, two 413 specification tests were passed. The two Hausman tests passed and confirmed the 414 appropriateness of using the AM specification instead of a basic FE or HT regressor. The 415 results of these tests are reported in

416 Table *6*.

417 As the AM estimator requires time invariant variables, we incorporate two additional 418 variables in the model: the size of countries and the mean value of patents. We choose the 419 size of countries because the recycling activity differs by country. This time invariant effect 420 has a strong influence on the waste recycling rate and on social behaviour. In addition, 421 knowledge accumulation will give rise to heterogeneous technology levels across countries. 422 Following Verdolini and Galeotti (2011), the average number of patents available (PAT) over 423 the period studied (2000-2012) is used as a proxy for the country-specific technological 424 changes. The results are reported in

- 425 Table 6 and suggest that, the overall picture is consistent with the results from both the
- 426 FE and 2SLS specifications.

	Hausman Taylor		
	No trend	With trend	
	(1)	(2)	
$\ln W$	0.0454***	0.0375***	
	(0.0081)	(0.0078)	
ln <i>K</i>	0.0905***	0.0503***	
	(0.0189)	(0.0189)	
ln <i>L</i>	0.304***	0.303***	
	(0.0473)	(0.0434)	
ln <i>M</i>	0.247***	0.250***	
	(0.0185)	(0.0188)	
ln <i>H</i>	0.0786***	0.0498***	
	(0.0137)	(0.0139)	
lnA	0.132***	0.123***	
	(0.0189)	(0.0188)	
Т		0.0127***	
		(0.0023)	
$t^2$		-0.000583***	
		(0.0001)	
Ep	0.113***	0.158***	
	(0.0430)	(0.0387)	
lnSize	-0.0916*	-0.0448	
	(0.0486)	(0.0440)	
Number of obs	260	260	
Hausman test 1 Chi <sup>2</sup> (4)	14.65**	9.44*	
Hausman test 2 Chi <sup>2</sup> (4)	0.07	1.06	
sargan-hansen	5.1890	5.1890	

Table 6 - Results based on Hausman Taylor Amemiya-MaCurdy method

	~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~		
428	Note: The table reports estimates	for the regressions	using Hausman
429	Taylor Amemiya-MaCurdy speci	fications. All test	statistics in the
430	table are robust to heteroscedastic	ity and arbitrary w	ithin-correlation.
431	* p<0.10, **p<0.05, ***p<0.01. V	alues in parenthesis	are the standard
432	errors		

#### 433 6. Transmissions Channels and Mechanisms

We found that positive relationship between waste recycling and economic performance 434 435 of a country. In this session, we try to identify the channels through which that waste recycling enhances the performance of an economy. Waste recycling may affect economic 436 437 outputs through different socio-economic channels. As the reduction of waste might have 438 positive consequences on environmental conditions but also might impact economic activities. Guided by economic theories (Pittel, 2010; Di Vita, 2001; Ayres, 1999; Hoel,
1978) this section explores the channels of R&D, labour and material supposed to be drivers
to the relationship between recycling and economic output.

The exiting literature highlights R&D as a main mechanism driving recycling through which the positive impact of waste recycling on economic output can be reinforced (Slade, 1980; Di Vita, 2001; Sidique et al., 2010). It seems acceptable that technology is needed in order to provide high quality secondary material able to concurrence raw material. Investment into R&D would provide recycled material sufficient economic and technic attractivity to rival virgin one. This variable is supposed to be an important and main mechanism driving the positive effect of recycling for economic development.

449 Chang (1997) states that the number of people working for waste recycling determines 450 the amount of waste that can be recycled and the benefit thereafter. Moreover, Di Vita (2001) 451 shows that labour is one of the main channels through which waste recycling impacts on economic performance. People are needed to sort waste, as sorting technology are not 452 453 sufficiently effective. Therefore, waste management and recycling industries are labour-454 intensive sectors. In a context of underutilization of the workforce - common in OECD -455 additional public investment in labour will be favourable to employment and growth -456 following the Keynesian multiplier theory. Labour could be one possible transmission 457 channel that can enhance the effect of waste recycling.

Another transmission channel could be material, because the economic performance of developed economies depends mainly on resource flows (Ayres, 1999). As shown by Hoel (1978), material recycling is bound by the raw material input into the economy. Di Vita (2001) confirms, on the one hand, that the amount of waste dedicated for recycling is bounded by the use of raw material in the economy. Additionally, waste recycling works as a

substitute for raw material as input for production on the other hand. We thus expect that this
channel – material - will lower the positive effect of waste recycling on economic outputs.

To identify the effects of the different channels, we extend the model by introducing interaction terms. We interact these supposed transmission channels with the waste recycled. The revised model to be estimated is as follows:

$$\ln(Y_{i,t}) = \beta_W ln(W_{i,t}) + \sum \beta_{WT} \ln(W_{i,t}) * \ln(T_{i,t}) + \beta_{iT} \boldsymbol{\mathcal{X}}_{i,t} + \varepsilon_{T_{i,t}}, \#(7)$$

where T represents the transmission channels, either R&D, labour or material input or both last one. The effect of waste recycling therefore includes both the direct and indirect effects through variable that are expected to explain the underlying mechanism of the relationship between waste and economic outputs. We can estimate the net effect of waste recycling on GDP when the channel reinforcement effect is captured using the formula:

$$\frac{\partial \ln(Y)}{\partial \ln(W)} = \beta_W + \sum_T \beta_{WT} E(T) , \#(8)$$

473 where T represents the transmission channels, either R&D, labour or material input.

The results are presented in Table 7 - Results of the estimations when considering R&D as a driver explaining the effect of MSW recycling on GDP and The two underlying channels being emerged from this study are labour and material flow. Thanks to these mechanisms, the total effect of waste recycling reaches over 0.06%. Overall, the estimates presented in **Erreur ! Référence non valide pour un signet.** are consistent with the results in Table 2. When both transmission channels are simultaneously included in the model, the direct effect of waste recycling on GDP is further enhanced.

Table 8. It is peculiar that empirical results do not highlight R&D as a mechanism
through which the positive impact of waste recycling on economic output can be reinforced.
Technology advancement is empirically not enhancing economic productivity through

recycling. When comparing results from Table 7 - Results of the estimations when considering R&D as a driver explaining the effect of MSW recycling on GDP and The two underlying channels being emerged from this study are labour and material flow. Thanks to these mechanisms, the total effect of waste recycling reaches over 0.06%. Overall, the estimates presented in **Erreur ! Référence non valide pour un signet.** are consistent with the results in Table 2. When both transmission channels are simultaneously included in the model, the direct effect of waste recycling on GDP is further enhanced.

Table 8, one can observe that R&D is actually refraining the positive impact of recycling on economic performance. The results can be explained by the already advance enough technology existing among the studied panel. OECD country secondary market is already mature, and no further R&D investment is needed in this industry. R&D investment in recycling prevent investments from beneficial sectors of recycling. Policy makers should redirect their investments into other drivers while considering recycling for sustaining their development.

				T: {R&D, La	bour, Material	]}			_
	FE	2SLS	FE	2SLS	FE	2SLS	FE	2SLS	_
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	
lnW	0.109	0.100**	0.233*	0.196***	0.597***	0.653***	0.402**	0.437***	
1-D & D*1-W	(0.1020)	(0.0497)	(0.1160)	(0.0574)	(0.1940)	(0.1630)	(0.1500)	(0.1430)	
IIIK&D*IIIW	-0.00782	-0.00679	-0.0240*	-0.0195***	0.00123	0.00866	-0.0307*	-0.0207**	
	(0.0126)	(0.0060)	(0.0137)	(0.0068)	(0.0138)	(0.0072)	(0.0159)	(0.0097)	
lnR&D	0.278	0.256**	0.592*	0.503***	0.0708	-0.0862	0.723**	0.517**	
	(0.2780)	(0.1290)	(0.2930)	(0.1430)	(0.3050)	(0.1540)	(0.3400)	(0.2030)	
lnL*lnW					0.0142	-0.00101	0.0446**	0.0255	
					(0.0226)	(0.0199)	(0.0206)	(0.0202)	
lnL	0.381**	0.378***	0.386**	0.376***	0.0863	0.366	-0.472	-0.117	
	(0.1340)	(0.0839)	(0.1400)	(0.0866)	(0.4680)	(0.3860)	(0.4290)	(0.3950)	
lnM*lnW					-0.0394	-0.0331*	-0.0413	-0.0319*	
					(0.0247)	(0.0183)	(0.0253)	(0.0168)	
lnM	0.235***	0.236***	0.240***	0.241***	1.081*	0.961**	1.096*	0.909***	
	(0.0258)	(0.0202)	(0.0285)	(0.0205)	(0.5250)	(0.3830)	(0.5300)	(0.3490)	
Net effect	0.0366*	0.0376**	0.0108	0.0157	0.0557*	0.06176***	0.0130	0.0221*	_
of W	(0.0870)	(0.0111)	(0.0177)	(0.1140)	(0.0283)	(0.0132)	(0.0237)	(0.0127)	_
Time trend	No	No	Yes	Yes	No	No	Yes	Yes	
Year FE	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	
Country FE	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	
Other control	Yes	Yes	Yes	Yes	Yes	Yes	Yes	Yes	
R2	0.899	0.891	0.917	0.909	0.903	0.894	0.92	0.913	
F	275.9	252.4	249.2	266	172	211.1	215.7	224.4	
Note: .	* p	o<0.10,	**p<0.05,	***p<0.01.	Values	in parent	thesis ar	e the	stand

498 Table 7 - Results of the estimations when considering R&D as a driver explaining the effect of MSW recycling on GDP

errors

The two underlying channels being emerged from this study are labour and material flow. Thanks to these mechanisms, the total effect of waste recycling reaches over 0.06%. Overall, the estimates presented in **Erreur ! Référence non valide pour un signet.** are consistent with the results in Table 2. When both transmission channels are simultaneously included in the model, the direct effect of waste recycling on GDP is further enhanced.

505 Table 8 - Results of the estimations of the effect of MSW recycling on GDP when 506 interactions are included

	FE	2SLS	FE	2SLS
	(1)	(2)	(3)	(4)
lnW	0.587***	0.769***	0.620***	0.746***
	(0.1440)	(0.1630)	(0.1350)	(0.1530)
lnL*lnW	0.0153	0.0418**	0.0186	0.0380**
	(0.0128)	(0.0174)	(0.0121)	(0.0169)
lnL	0.0685	-0.447	-0.0211	-0.398
	(0.2580)	(0.3470)	(0.2430)	(0.3310)
lnM*lnW	-0.0393***	-0.0694***	-0.0439***	-0.0656***
	(0.0137)	(0.0170)	(0.0129)	(0.0159)
ln <i>M</i>	1.076***	1.708***	1.178***	1.632***
	(0.2890)	(0.3570)	(0.2720)	(0.3320)
Net effect of W	0.05445***	0.06179***	0.048201***	0.05335***
	(0.0087)	(0.0103)	(0.0082)	(0.0086)
Time trend	No	No	Yes	Yes
Year FE	Yes	Yes	Yes	Yes
Country FE	Yes	Yes	Yes	Yes
Other controls	Yes	Yes	Yes	Yes
$\mathbb{R}^2$	0.896	0.893	0.908	0.907
F	282	229.5	259.6	220.7

Note: \* p<0.10, \*\*p<0.05, \*\*\*p<0.01. Values in parenthesis are the standard errors

508 Our results on these two transmission mechanisms are consistent with the literature 509 (Pittel, 2010; Di Vita, 2001; Ayres, 1999; Chang, 1997; Hoel, 1978). We therefore show the 510 evidence of the Keynesian multiplier effects: injecting additional public investment in 511 workforce and material efficiency towards the recycling sector, will further benefit and 512 sustain the economic development.

513 On the one hand, the indirect effect of material use is negative for the reasons explain 514 above regarding the substitution effect between waste and material (Hoel, 1978). Indeed, to a 515 certain level, recycled materials can substitute for virgin resources. On the other hand, the 516 overall effect of material on economic outputs is further enhanced while interacting with 517 recycled waste. The rationale is that recycling is bounded by the injection of raw material into 518 the economy. As the literature emphasises the material input limits the recycling rate. Our 519 findings provide a new direction for exploring the positive effects of recycling. Our results 520 are in line with the existing literature regarding the substitutional equivalence between 521 recycled and virgin materials (Calcott and Walls, 2005). Additionally, we demonstrate the 522 strong connection between virgin material input and recycling possibilities. As mentioned by 523 Pittel et al. (2010), there is a significant lack of investment in the recycling sector. Therefore, 524 increasing investment in recycling could decrease the dependency that economic output 525 currently has on raw material input. This would allow developed countries to move towards a 526 dematerialisation of the economy.

With regard to labour, the indirect effect is positive while the overall effect is negative. Our results are consistent with the literature on the role of labour as a driver for the recycling industry. Increasing recycling requires labour, which is taken from the labour pool available and directed towards the recycling sector. The total direct effect is positive and confirms the role of labour as a transmission channel that can enhance the effect of waste recycling. Figure 8 illustrates the overall marginal effects of waste recycling across countries, when presenting the underlying mechanism positively driving the relationship between recycling and economic outputs.



536 Figure 8 – Overall marginal effect of waste recycling on economic performance per country

535

537 The first observation is that the effect is positive for all countries except the United States. 538 Furthermore, we find that the marginal effect is heterogeneous across countries. Possible 539 explanations for this result are the variation in levels of training of the labour force in 540 different countries, cultural differences in terms of working time, the import and export of 541 materials, and the structural difference in terms of waste management. The United States is 542 the only country among the panel studied where these interaction effects of labour and 543 material inputs with waste recycled are negative. Indeed, the United States has more labour involved in the extraction industry than in recycling. Additionally, a significant portion of the 544 545 United States' economic output is based on raw material trade. The United States has 546 depletion subsidies for extraction as well as reduction taxes for the transportation of raw

materials but not for scraps, which unbalances the economy of recycled material (Slade,1980). If this country is excluded from the analysis, the results are similar (see Appendix C).

When the channels are considered individually, the results are less significant. This lack of significance can be explained by the heterogeneity of the effects of the transmission channels among countries. With regard to labour, the difference in working policies can explain such disparities. With regard to materials, the unbalanced economic importance of imports and exports of materials can explain the disparities.

In fact, the development of the recycling industry brings in new business and job opportunities for low educated workforce. It is of particular interest for mature economies, as it reactivates both their educated -and less educated- labour pool. It also helps to decrease the unemployment rate in OECD countries, which already have a workforce with high environmental awareness. Moreover, the economy can grow by using recycled materials instead of extracting new ones, and thus cut the use of scarce natural resources.

#### 560 **7. Conclusion**

This paper provides the first empirical study on the positive effects of waste recycling on 561 562 the economic performance of developed countries. We approach the question from a 563 macroeconomic aspect and provide empirical evidence to support the linkage between 564 economic growth and waste recycling. We quantify the effect of waste recycling on data from 565 OECD over 13 years (2000-2012). This novel paper complements the scarce empirical 566 literature where the positive impact of waste recycling has already been highlighted at 567 microeconomic level. We find positive and statistically significant effects of waste recycling 568 on the economy. Our results show that an increase of 1% in recycled waste leads to an 569 increase in GDP by up to 0.06%. This effect is also stronger at the 10% level when considering per capita ratio. Our results are robust to a series of sensitivity checks, controlling 570

571 for reverse causality, endogeneity, and other econometric biases. Our results provide robust 572 evidence confirming the existing theoretical models that link growth and recycling.

573 This study provides strong support for establishing environment and resource recovery 574 policies, at least for OECD countries. These findings could be used in developed countries to 575 assist policy makers to enforce recycling policy. Our paper highlights that waste 576 management, if well planned, can be a catalyst for a stagnating economy. Using alternative 577 measures of waste recycling, our results suggest that recycling policy at household level 578 results in stronger economic outcomes, and the government should promote higher market 579 penetration for the waste recycling industry. This paper foregrounds waste management as a 580 new growth point for stagnating economies.

581 In addition, we highlight the underlying drivers of the relationship between recycling and 582 economic outcomes. Our empiric findings show that R&D is not a mechanism enhancing 583 economic performance through recycling, contradicting expectations. The benefits of waste 584 recycling are rather drive through two channels: labour and material use. Investments should 585 be intensified in the recycling sector to enhance its positive effect on economic performance. 586 Such investments would decrease the dependency of economic output on raw material input 587 and allow a transition to a dematerialisation of the economy. Furthermore, it will reactivate 588 their labour pool and decrease the unemployment rate. Besides, governments could consider 589 fostering labour in the recycling industry, as it will not only enhance environmental 590 management but also bring in benefits to overall economic performance.

591 Managing waste through recycling could be an effective way to reduce the cost of 592 environmental policies, as it lowers the greenhouse gas emissions and prevents environmental 593 degradation due to excess resource extraction and consumption. As such, recycling can be 594 considered a global and competitive strategy for climate change mitigation. These 595 conclusions might be of a special interest for policy makers devoting to circular economy.

A limitation of the study is the use of country as the unit of analysis. Waste management policies and practices are mostly locally governed and can vary significantly within a country. A microeconomic approach might be more appropriate for studying the local responsibilities of waste recycling among OECD countries. However, obtaining precise data on waste management at the local level is a significant challenge. Additionally, the recycling sector is currently experiencing considerable change. It is of a high interest to compare the results with similar analyses that will be done in the next decennial.

#### 603 Appendix A

_		ln <i>K</i>	lnL	$\ln M$	ln <i>H</i>	lnA	residual
-	ln <i>K</i>	1					
	ln <i>L</i>	0.563***	1				
	ln <i>M</i>	0.499***	0.947***	1			
	ln <i>H</i>	0.332***	0.432***	0.454***	1		
	lnA	0.571***	0.845***	0.827***	0.729***	1	
	residual	-0.317***	0.205***	0.234***	0.414***	0.420***	1

This appendix compiles the different tests carried out on the data.

605		
	605	
	00.0	

Table A 1 - Correlation matrix between the residuals and the control variables

607through this approach is then extracted to study its correlation with the control variable. Similar results are608obtained for the correlation between the control variables and the residuals estimated for equation 3 through the609same approach. \* p<0.10, \*\*p<0.05, \*\*\*p<0.01</td>

610

611

### Table A 2 - Granger Causality test

	lnW->lnGDP	lnGDP->lnW
N	12	12
adj. R <sup>2</sup>	0.907	0.569
wbar	2.5959	8.1228
zbar	5.0468***	13.6910***
zbart	4.4090***	2.5477**

<sup>612</sup> Note: The test applied is the one suggested by Dumitrescu and Hurlin (2012)

614 H1: X does Granger-cause Y for at least one panel var (id)

615 \* p<0.10, \*\*p<0.05, \*\*\*p<0.01.

H0: X does not Granger-cause Y.

	Water&EPS			EPS			Water		
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)
·								-	·
Water	-1.294***	-1.352***	0.0968	-	-		-1.288***	1.354***	1.035***
	(0.2260)	(0.2250)	(0.181)	-	-		(0.2598)	(0.2502)	(0.1670)
EPS	0.0663**	0.00668	0.287***	0.0636*	0.0218	0.297***	-	-	
	(0.0351)	(0.0436)	(2.9230)	(0.0486)	(0.0474)	(0.0263)	-	-	
Time trend	по	yes	no	no	yes	no	no	yes	no
Add. Control	yes	yes	no	yes	yes	no	yes	yes	no
KP Wald F stat.									
(weak id)	13.13	15.29		1.71	0.22		24.57	29.27	
KP Wald stat.									
(under id)	27.04***	31.77***		1.76	0.21		25.20***	30.28***	
KP LM stat.									
(under id)	18.092***	18.55***		1.66	0.22		17.75***	18.00***	
partial R <sup>2</sup>	0.1339	0.136	0.292	0.0122	0.0091	0.294	0.1206	0.1359	0.066

### 617 Table A 3 – Statistical test results of the first step of 2SLS IV approach

Note: The results are estimated with 2SLS. The dependant variable is  $\ln W$ . Country and year fixed effects are included. Standard errors are clustered and reported into brackets. All test statistics in the table are robust to heteroscedasticity and arbitrary within-correlation. KP holds for the Kleibergen and Paap (2006) tests. They account for underestimation and weak identification instruments (Bazzi and Clemens, 2013). \* p<0.10, \*\*p<0.05, \*\*\*p<0.01.

	Fisher unit-root test						
	For panel unit root	Unit-root test based on augmented					
	using an augmented		ckey Fuller to	augmented			
	Dickey–Fuller test		$(0 \ln z)$	-515			
	(0 lag)	(U lag)					
		Inverse	Inverse	Modified			
	Inverse Chi <sup>2</sup> (40)	normal	logit t	Inverse			
		normar	(104)	Chi <sup>2</sup> (40)			
lnGDP	89.8151***	-5.1562***	-5.0579***	5.5695**			
lnW	80.2782***	-3.6677***	-3.8000***	4.5032**			
ln <i>K</i>	111.1646***	-5.6444***	-6.1568***	7.9564**			
ln <i>L</i>	70.0210***	-2.9081***	-2.8754***	3.3564**			
ln <i>H</i>	83.1695***	-4.1707***	-4.2858***	4.8265**			
ln <i>M</i>	96.9172***	-5.3537***	-5.1814***	5.5689**			
lnA	112.6701***	-3.5101***	-5.0158***	8.1248**			
EPS	98.4427***	-5.1716***	-5.3176***	6.5341**			
Water	113.2944***	-4.2669***	-4.8650***	8.1946**			
Note: Inverse chi-squared (40) P							
Inverse norm	al Z						
Inverse logit	t (104) L*						
Modified inv. chi-squared Pm							

# 624 Table A 4 - Augmented Dickey-Fuller stationary test results on the panel data

626	Inverse normal Z
627	Inverse logit t (104) L*
628	Modified inv. chi-squared Pm
629	H0: All panels contain unit roots
630	Ha: At least one panel is stationary
631	Standard errors are clustered and reported into brackets. * p<0.10, **p<0.05, ***p<0.01
632	

**Appendix B** This appendix summarizes the definitions of the variables used in this paper.

Table B 1 - Data Sources and definitions of variables of interest

Variable	Definition	Source
Pop	Total population	UN (2017)
GDP	Gross domestic product (indicator)	OECD (2018a)
W	Municipal solid waste amount designated for recycling and energy recovery operations	OECD (2018b)
Кр	Private capital stock (current cost)	IMF (2017)
Kg	General government capital stock	-
L	Employment rates are defined as a measure of the extent to which available labour resources (people available to work) are being used.	OECD (2018c)
М	DMI measures the direct input of materials used in production and consumption activities (excluding water flows). It is calculated as domestic extraction plus imports (in economy-wide material flow accounting).	UNEP (2016)
Н	Government expenditure for tertiary students	World Bank (2018)
R&N	Gross domestic expenditure on R-D	OECD (2018d)
Water	Waste water treatment indicator presents sewage treatment connection rates	OECD (2018e)
EPS	Environmental Policy Stringency Index	Botta and Kozluk (2014)
РАТ	Number of patents available by country under the classification of International Patent Classification (IPC), referenced by inventor(s)'s country(ies) of residence	OECD (2018d)
Country Size	Size of countries. International territories acquired during colonial periods were not considered.	Combined by authors
DMC	DMC measures the total amount of materials directly used by an economy and is defined as the annual quantity of raw materials extracted from the domestic territory, plus all physical imports minus all physical exports.	UNEP (2016)

## Appendix C

This appendix details the results obtained for the transmission channels when the country "United State" is dropped.





transmission channels are considered (US dropped)

	T: {Labour, Material} - US dropped					
	FE	2SLS	FE	2SLS		
	(1a)	(1b)	(2a)	(2b)		
$\ln W$	0.588***	0.354*	0.613***	0.883**		
	(0.1480)	(0.377)	(0.141)	(0.345)		
lnL*lnW	-0.00475	-0.0164	0.00109	-0.0107		
	(0.0143)	(0.0170)	(0.0136)	(0.0162)		
ln <i>L</i>	0.456	0.681**	0.320	0.553*		
	(0.2860)	(0.3270)	(0.2720)	(0.3080)		
lnM*ln₩	-0.023	-0.00205	-0.0293**	-0.0334		
	(0.0153)	(0.0259)	(0.0145)	(0.0229)		
ln <i>M</i>	0.755**	0.314	0.887***	0.985**		
	(0.3190)	(0.5450)	(0.3030)	(0.4830)		
Net effect of W	0.05723***	0.05089***	0.05285***	0.05307***		
	(0.0086)	(0.0082)	(0.0092)	(0.0100)		
Time trend	No	No	Yes	Yes		
Country FE	Yes	Yes	Yes	Yes		
$R^2$	0.901	0.899	0.912	0.909		
F	283.3	256.35	240.6	191.5		

Table C 1 - Results of the estimations of the effects of MSW recycling on GDP when interactions are included and United States dropped

Note: The results are estimated with 2SLS and FE. FE: fixed-effect ordinary least squares and 2SLS: two-stage least squares with external instruments. The dependent variable is ln*GDP*. The estimates for the transmission channels variables are consistent with the results in The two underlying channels being emerged from this study are labour and material flow. Thanks to these mechanisms, the total effect of waste recycling reaches over 0.06%. Overall, the estimates presented in **Erreur ! Référence non valide pour un signet.** are consistent with the results in Table 2. When both transmission channels are

simultaneously included in the model, the direct effect of waste recycling on GDP is further enhanced.

Table 8. The estimates for all the other variables are positive, significant at least at the 5% level and consistent with the results in Table 2 The Hausman test was conducted to confirm the validity of the choice for FE. For the 2SLS, EPS and Water are used jointly as external instrument variables. Standard errors are clustered and reported in brackets. All test statistics in the table are robust to heteroscedasticity and arbitrary within-correlation. \* p<0.10, \*\*p<0.05, \*\*\*p<0.01

#### References

- Abbott, A., Nandeibam, S., & O'Shea, L. (2017). The displacement effect of convenience: the case of recycling. Ecological Economics, 136, 159-168.
- Acemoglu, D., Johnson, S., & Robinson, J. A. (2001). The colonial origins of comparative development: An empirical investigation. American economic review, 91(5), 1369-1401.
- Acuff, K., & Kaffine, D. T. (2013). Greenhouse gas emissions, waste and recycling policy.Journal of Environmental Economics and Management, 65(1), 74-86.
- Aghion, P., & Howitt, P. (1992). The Schumpeterian approach to technical change and growth. In Economic Growth in the World Economy: Symposium (Vol. 1993).
- Albrizio, S., Kozluk, T., & Zipperer, V. (2017). Environmental policies and productivity growth: Evidence across industries and firms. Journal of Environmental Economics and Management, 81, 209-226.
- Almer, C., & Winkler, R. (2017). Analyzing the effectiveness of international environmental policies: The case of the Kyoto Protocol. Journal of Environmental Economics and Management, 82, 125-151.
- Alpízar, F., & Gsottbauer, E. (2015). Reputation and household recycling practices: field experiments in Costa Rica. Ecological Economics, 120, 366-375.

- Andersen, F. M., Larsen, H., Skovgaard, M., Moll, S., & Isoard, S. (2007). A European model for waste and material flows. Resources, conservation and recycling, 49 (4), 421–435.
- Ayres, R. U. (1999). The second law, the fourth law, recycling and limits to growth. Ecological Economics, 29(3), 473-483.
- Barro, R. J. (1996). Determinants of economic growth: a cross-country empirical study (No. w5698). National Bureau of Economic Research.
- Bazzi, S., & Clemens, M. A. (2013). Blunt instruments: Avoiding common pitfalls in identifying the causes of economic growth. American Economic Journal: Macroeconomics, 5(2), 152-86.
- Botta, E., & Koźluk, T. (2014). Measuring environmental policy stringency in OECD countries.
- Bretschger, L., Lechthaler, F., Rausch, S., & Zhang, L. (2017). Knowledge diffusion, endogenous growth, and the costs of global climate policy. European Economic Review, 93, 47-72.
- Burtraw, D., Linn, J., Palmer, K., & Paul, A. (2014). The costs and consequences of clean air act regulation of CO2 from power plants. American Economic Review, 104(5), 557-62.
- Calcott, P., & Walls, M. (2005). Waste, recycling, and "Design for Environment": Roles for markets and policy instruments. Resource and Energy Economics, 27(4), 287-305.
- Chang, N.-B., & Lin, Y. (1997). An analysis of recycling impacts on solid waste generation by time series intervention modeling. Resources, Conservation and Recycling, 19 (3), 165-186.
- Ciacci, L., Passarini, F., & Vassura, I. (2017). The European PVC cycle: In-use stock and flows. Resources, Conservation and Recycling, 123 ,108–116.

- D'Amato, A., Mancinelli, S., & Zoli, M. (2016). Complementarity vs substitutability in waste management behaviors. Ecological Economics, 123, 84-94.
- Dickey, D. A., & Fuller, W. A. (1979). Distribution of the estimators for autoregressive time series with a unit root. Journal of the American statistical association, 74(366a), 427-431.
- Di Vita, G. (2001). Technological change, growth and waste recycling. Energy Economics, 23(5), 549-567.
- European Commission. (2010). Breaking the link between economic growth and waste generation - DG Environment News Alert Service. http://ec.europa.eu/environment/integration/research/newsalert/-pdf/203na2 en.pdf. ([Online; accessed 5-December-2017])
- European Commission. (2015). Closing the loop an EU action plan for the circular economy. Communication from the Commission to the European Parliament, the Council, the European economic and Social Committee and the Committee of the Regions. Brussels: COM .
- Fagnart, J. F., & Germain, M. (2011). Quantitative versus qualitative growth with recyclable resource. Ecological Economics, 70(5), 929-941.
- FREC, Ecologique-solidaire.gouv.fr. (2018). Feuille de route économie circulaire (FREC). [online] Available at: https://www.ecologiquesolidaire.gouv.fr/sites/default/files/Feuille-de-route-Economie-circulaire-50-mesurespour-economie-100-circulaire.pdf [Accessed 9 Jul. 2018].
- Gilli, M., Nicolli, F., & Farinelli, P. (2018). Behavioural attitudes towards waste prevention and recycling. Ecological Economics, 154, 294-305.
- Granger, C. W. (1969). Investigating causal relations by econometric models and crossspectral methods. Econometrica: Journal of the Econometric Society, 424-438.

- Gupta, S., Kangur, A., Papageorgiou, C., & Wane, A. (2014). Efficiency-adjusted public capital and growth. World Development, 57, 164-178.
- Hamilton, S. F., Sproul, T. W., Sunding, D., & Zilberman, D. (2013). Environmental policy with collective waste disposal. Journal of Environmental Economics and Management, 66(2), 337-346.
- Harford, J. D. (1976). Adjustment costs and optimal waste treatment. Journal of Environmental Economics and Management, 3(3), 215-225.
- Hausman, J. A., & Taylor, W. E. (1981). Panel data and unobservable individual effects. Econometrica, 49(6), 1377-1398.
- Helpman, E. (1992). Endogenous macroeconomic growth theory. European Economic Review, 36(2-3), 237-267.
- Hoel, M. (1978). Resource extraction and recycling with environmental costs. Journal of Environmental Economics and Management, 5(3), 220-235.
- IMF (2017). International Monetary Fund. Investment and Capital Stock Dataset, 1960-2015 http://www.imf.org/external/np/fad/publicinvestment/data/info122216.pdf
- IPCC (2014), Blanco G, Gerlagh R, Suh S, Barrett J, de Coninck HC, Morejon CFD, Mathur R, Nakicenovic N, *et al.* (2014). Chapter 5 Drivers, trends and mitigation. In: Climate Change 2014: Mitigation of Climate Change. IPCC Working Group III Contribution to AR5. Cambridge University Press.
- IPCC. (2007). Report of the nineteenth session of the Intergovernmental Panel on Climate Change (IPCC) Geneva,
- IPCC (2006), IPCC guidelines for national greenhouse gas inventories volume 5—waste H. Eggleston, L. Buendia, K. Miwa, T. Ngara, K. Tanabe (Eds.), The National Greenhouse Gas Inventories Programme, the Intergovernmental Panel on Climate Change. Hayama, Kanagawa, Japan (2006), pp. 6.1-6.28

- Jenkins, R. R., Martinez, S. A., Palmer, K., & Podolsky, M. J. (2003). The determinants of household recycling: a material-specific analysis of recycling program features and unit pricing. Journal of Environmental Economics and Management, 45(2), 294-318.
- Johnstone, N., & Labonne, J. (2004). Generation of household solid waste in OECD countries: an empirical analysis using macroeconomic data. Land Economics, 80 (4), 529-538.
- Kamps, C. (2006). New estimates of government net capital stocks for 22 OECD countries, 1960–2001. IMF staff papers, 53(1), 120-150.
- Kaza, S.; Yao, L. C.; Bhada-Tata, P.; Van Woerden, F. (2018). What a Waste 2.0 : A Global Snapshot of Solid Waste Management to 2050. Urban Development. Washington, DC:
  World Bank. © World Bank. https://openknowledge.worldbank.org/handle/10986/30317 License: CC BY 3.0 IGO."
- Kellenberg, D. (2012). Trading wastes. Journal of Environmental Economics and Management, 64(1), 68-87.
- Kinnaman, T. C., Shinkuma, T., & Yamamoto, M. (2014). The socially optimal recycling rate: evidence from Japan. Journal of Environmental Economics and Management, 68(1), 54-70.
- Kinnaman, T. C., & Fullerton, D. (2000). Garbage and recycling with endogenous local policy. Journal of Urban Economics, 48 (3), 419–442.
- Kormendi, R. C., & Meguire, P. G. (1985). Macroeconomic determinants of growth: crosscountry evidence. Journal of Monetary Economics, 16(2), 141-163.
- Lacy, P., & Rutqvist, J. (2016). Waste to wealth: The circular economy advantage. Springer.
- Mavropoulos, A. (2010). Waste management 2030+. Waste management world, 11(2).
- Mazzanti, M., & Zoboli, R. (2008). Waste generation, waste disposal and policy effectiveness: Evidence on decoupling from the European union. Resources, Conservation and Recycling, 52 (10), 1221-1234.

- Mazzanti, M., & Zoboli, R. (2009). Municipal waste Kuznets curves: evidence on socioeconomic drivers and policy effectiveness from the EU. Environmental and Resource Economics, 44(2), 203.
- METI (2001). Law for promotion of effective utilization of resources. Ministry of Economy, Trade and Industry. METI, Japan. Retrieved from <u>http://www.meti.go.jp/policy/recycle/main/english/-pamphlets/pdf/cReEffectLe</u> <u>2006.pdf</u>
- Milligan, B., & O'Keeffe, M. (2019). Global governance of resources and implications for resource efficiency in Europe. Ecological Economics.
- OECD (2018a), Gross domestic product (GDP) (indicator). doi: 10.1787/dc2f7aec-en

OECD (2018b), Municipal waste (indicator). doi: 10.1787/89d5679a-en

OECD (2018c), Employment rate (indicator). doi: 10.1787/1de68a9b-en

- OECD (2018d). "Patents by main technology and by International Patent Classification (IPC)", OECD Patent Statistics (database). doi.org/10.1787/data-00508-en
- OECD (2018e), Waste water treatment (indicator). doi: 10.1787/ef27a39d-en (Accessed on 18 December 2018)
- Pittel, K., Amigues, J. P., & Kuhn, T. (2010). Recycling under a material balance constraint. Resource and Energy Economics, 32(3), 379-394.
- Romer, P. M. (1990). Endogenous technological change. Journal of political Economy, 98(5, Part 2), S71-S102.
- Schulze, W. D. (1974). The optimal use of non-renewable resources: The theory of extraction. Journal of Environmental Economics and Management, 1(1), 53-73.
- Shinkuma, T. (2003). On the second-best policy of household's waste recycling. Environmental and Resource Economics, 24(1), 77-95.

- Sidique, S. F., Joshi, S. V., & Lupi, F. (2010). Factors influencing the rate of recycling: An analysis of Minnesota counties. Resources, Conservation and Recycling, 54 (4), 242– 249.
- Slade, M. E. (1980). An econometric model of the US secondary copper industry: Recycling versus disposal. Journal of Environmental Economics and Management, 7(2), 123-141.
- Smith, V. L. (1972). Dynamics of waste accumulation: disposal versus recycling. The Quarterly Journal of Economics, 86(4), 600-616.
- Stock, J.H., & Yogo, M. (2005). Asymptotic distributions of instrumental variables statistics with many instruments (Vol. 6). Chapter.
- Ten Brink, P., Schweitzer, J.-P., Watkins, E., Janssens, C., De Smet, M., Leslie, H., & Galgani, F. (2018). Circular economy measures to keep plastics and their value in the economy, avoid waste and reduce marine litter (Tech. Rep.). Economics Discussion Papers.
- UN (2017). United Nations, Department of Economic and Social Affairs, Population Division. World Population Prospects: The 2017 Revision, Data Booklet. ST/ESA/SER.A/401.
- Verdolini, E., & Galeotti, M. (2011). At home and abroad: An empirical analysis of innovation and diffusion in energy technologies. Journal of Environmental Economics and Management, 61(2), 119-134.
- World Bank (2018). The World Bank, World Development Indicators. Current education expenditure, primary (% of total expenditure in primary public institutions), Median method [United Nations Educational, Scientific, and Cultural Organization Institute for Statistics].
   Retrieved from https://data.worldbank.org/indicator/SE.XPD.CPRM.ZS?view=chart
- UNEP (2016). Global Material Flows and Resource Productivity. An Assessment Study of the UNEP International Resource Panel. H. Schandl, M. Fischer-Kowalski, J. West, S.

Giljum, M. Dittrich, N. Eisenmenger, A. Geschke, M. Lieber, H. P. Wieland, A. Schaffartzik, F. Krausmann, S. Gierlinger, K. Hosking, M. Lenzen, H. Tanikawa, A. Miatto, and T. Fishman. Paris, United Nations Environment Programme.