

1 Do Forest-Management Plans and FSC certification reduce deforestation in the Congo Basin?

2 Highlights

- 3 - The adoption of sustainable forest-management practices that are FSC-certified or with an
4 accepted forest-management plan is associated with less deforestation in the Congo Basin
- 5 - This lower deforestation takes at least five years to come about in concessions with a forest-
6 management plan
- 7 - The fall in deforestation is more pronounced around communities located within or nearby to
8 logging concessions and close to sites of previous deforestation.

9 **Abstract:** To allow for the production of timber while preserving conservation values, forestry
10 regulations in the Congo Basin have made Forest Management Plans (FMPs) mandatory in logging
11 concessions. This paper uses original high-resolution maps of forest-cover changes and official records
12 on the activities of logging concessions to analyze the impact of FMPs on deforestation in this region.
13 We apply quasi-experimental and difference-in-difference approaches to evaluate the change in
14 deforestation in concessions that implemented an FMP. We find that deforestation fell by 74%
15 between 2000 and 2010 in concessions with an FMP. Building on a theory of change, further analyses
16 revealed that this lower deforestation takes at least five years to come about, and is highest around
17 communities located in and nearby logging concessions and in areas close to previous deforestation.
18 These findings suggest that FMPs reduce deforestation by allowing concessions to rotate cycles of
19 timber extraction, thereby avoiding the (over-) exploitation of areas that were previously logged, and
20 by the better regulation of access to concessions by closing former logging roads to limit illegal
21 activities such as slash and burn agriculture, hunting and the illegal harvest of timber or fuelwood.

22 **Keyword:** forest management plan; FSC certification; deforestation; quasi-experimental matching;
23 causal mechanisms; Congo basin.

24 1 Introduction

25 About 400 million hectares of natural tropical forest are devoted to timber production (Blaser, 2011).
26 Ensuring the sustainable exploitation of these forests is a crucial challenge, as they are a key factor for
27 biodiversity, carbon sequestration and the global climate. In the Congo Basin, the second-largest
28 tropical forest after the Amazon, with an area of about 178 million ha of dense humid forests (Mayaux
29 et al., 2013), almost one third of forests are productive in terms of logging exploitation. National
30 forestry regulations have made Forest Management Plans (FMPs) mandatory in logging concessions to
31 ensure their sustainable exploitation. The FMP must ensure sustainable forest management, that is
32 timber production that limits deforestation and guarantees the preservation of forest resources,
33 biodiversity and ecosystem services, while contributing to local socio-economic development (Nasi et
34 al., 2012).

35 For this reason, and because of the extent of forest areas covered, FMPs are often considered as a
36 major contribution to tropical forest conservation worldwide, and have been supported by
37 international organizations and NGOs (Clark et al., 2009; Lambin et al., 2014). However, there is
38 relatively scant empirical work on their effect on deforestation in logging concessions. Cerutti et al.
39 (2017) showed that FMPs in Cameroon between 1998 and 2009 effectively reduced carbon emissions
40 from logging operations due to the reduced volumes of timber harvested, as imposed by the FMP,
41 while presenting logging companies with acceptable financial trade-offs. On the contrary, Brandt et al.
42 (2016) found that FMP concessions in the Congo, compared to otherwise similar concessions without,
43 were associated with greater deforestation. Further analyses suggested that, greater timber
44 production driven by increased foreign capital and international demand contributed to greater
45 deforestation in the six concessions with FMPs in the Congo (Brandt et al., 2014, 2016). This led to a
46 controversy between Karsenty et al. (2017) and Brandt et al. (2018), emphasizing the need for more
47 empirical work to understand whether and under which conditions FMPs affect deforestation.

48 While there is a paucity of work on the effects of FMPs, relatively more attention has been given to
49 Forest Steward Council (FSC) certification: this is a voluntary market-based approach to enhance
50 sustainable forest management. As halting tropical deforestation remains a central FSC objective,
51 within a wide range of issues covered by FSC standards, a number of empirical contributions have
52 looked at the impact of FSC certification on deforestation. The results here are also mixed and context-
53 dependent. Some work on Cameroon (Panlasigui et al. 2018), Mexico (Blackman et al. 2018), and Brazil,
54 Gabon and Indonesia (Rana and Sills 2017) has shown that FSC certification reduced deforestation in
55 most certified logging concessions, but that the estimated effects were rarely statistically different
56 from zero and varied over time, thus providing inconclusive evidence of the deforestation impact of
57 FSC. Miteva et al. (2015) showed that FSC certification in Indonesia reduced deforestation and
58 improved household welfare. In Chile, Heilmayr and Lambin (2016) compared the deforestation
59 impacts of three different non-State market-driven governance regimes, among which FSC
60 certification: they showed that FSC certification effectively reduced deforestation, and was more
61 effective than the other measures tested, which were more industry-friendly.

62 Overall, the impact of the adoption of sustainable forest-management practices on deforestation in
63 the Congo Basin remains an active research area. The results from similar policy interventions in Asia
64 and South America suggest that the results are context-dependent and can therefore not be directly
65 transposed. As reducing deforestation in low-income countries is arguably one of the most cost-
66 effective ways of reducing global CO₂ emissions (Stern, 2006, and Barker et al., 2007), this paper aims
67 to evaluate the change in forest cover following the implementation of an FMP or FSC certification in
68 the Congo Basin, and to establish the underlying mechanisms explaining whether and how these work
69 (Baylis et al., 2016; Miteva et al., 2012).

70 To provide an empirical estimate of the impact of FMPs in the Congo Basin, we use original high-
71 resolution maps of changes in forest cover in Cameroon, Congo, Gabon and the Central African
72 Republic (CAR) over the 1990-2000 and 2000-2010 periods. The geographic area does not include

73 forest-cover changes in the Democratic Republic of Congo, where FMPs were initiated later. The
74 deforestation maps are complemented with relevant detailed information on the location and extent
75 of logging concessions, including the timing of the official approval of their FMP and FSC certification.
76 As the selection into FMP adoption is not random, we use quasi-experimental methods whereby the
77 logging concessions that adopted FMP are compared to logging concessions that did not adopt an FMP
78 but had otherwise similar observable characteristics that are known to affect deforestation.

79 This approach will likely produce unbiased estimates of the effect of FMPs in the study areas for at
80 least two reasons. First, since the 1990's, Cameroon, Congo, CAR and Gabon have all implemented
81 reforms aimed at encouraging logging companies to adopt FMPs (Karsenty, 2006). FMP were then
82 gradually implemented in the 2000s, and by 2010 almost one-third of the concessions in the study area
83 had an accepted FMP. FSC certification is more recent in the region, starting only in 2005. Given the
84 staggered rollout of reforms promoting FMP adoption in the region, it is likely that we will find
85 otherwise-similar concessions with and without FMPs, which is a key requirement for unbiased quasi-
86 experimental analysis. Second, even though national policies aiming to increase FMP adoption have
87 been discussed since the 1990s, the first logging concessions with FMPs appeared in the early 2000s
88 in the Congo Basin. Since we can also measure deforestation between 1990 and 2000, we fine-tune
89 our estimates of the FMP impact on logging concessions by correcting for pre-existing differences in
90 deforestation rates between early and late FMP adopters in the Congo Basin. Last, we test the
91 robustness of the results and replicate our analysis in data from the widely-used Global Forest Change
92 (GFC) dataset (Hansen et al., 2013) over the 2000-2010 period. By doing so, we add to existing
93 empirical work by considering the Congo Basin. As we cover a larger sample of logging concessions, we
94 avoid the limitations of analyses based on smaller samples.

95 The remainder of the paper is organized as follows. In Section 2 we present background information
96 on forest-management plans and the theoretical framework behind their potential deforestation
97 effects in the Congo Basin. Section 3 then describes the main datasets used, and Section 4 outlines the

98 empirical strategy used to explore the causal impact of FMPs on deforestation. Section 5 presents the
99 main results and their robustness and limitations, and explores the channels underlying the link
100 between FMPs and deforestation. Last, Section 6 discusses the implications of our work and offers
101 some concluding observations.

102 2 Background and theoretical framework

103 In the Congo Basin, most forested areas are State-owned, and exploitation permits are granted to
104 private logging companies for long periods (up to 100 years) under concession regimes, providing long-
105 term resource-extraction rights in exchange for a stream of revenues (Agrawal et al., 2008). In this
106 context FMPs are a tool for sustainable forest management, combining timber production, local
107 development and conservation values in the Congo Basin.

108 2.1 Forest-Management Plans in the Congo Basin

109 FMPs in a concession involve a range of environmental and social issues. They are based on forest
110 inventories describing the distribution of trees species and their characteristics. Based on ecological
111 and social studies (e.g., on fauna and the forest uses of local communities), these inventories allow us
112 to divide each concession into “management series” areas according to the use of forest resources.
113 Among these, the “production”, “conservation” and “community management” series respectively
114 refer to: wood exploitation; the preservation of biodiversity, seed trees and the most vulnerable areas
115 (with buffer zones on steep slopes, riversides etc.); and last local-community development. These
116 community-management series are located around settlements and agricultural areas, and aim to
117 ensure the coexistence of different forest uses in order to guarantee the land rights of local populations
118 and encourage local communities to carry out sustainable natural-resource management, in particular
119 regarding hunting and agriculture (ATIBT, 2007; Nkeoua, 2003). The production series are divided into
120 "annual cutting areas (ACA)", for which the FMP presents a detailed plan for selective logging over a
121 specific time period. This plan aims to optimize the exploitation of timber, while ensuring the
122 regeneration of forest species in order to guarantee the viability of the next logging cycle (the usual,

123 rotation time is between 25 and 30 years). In addition, FMPs recommend reduced-impact logging (RIL)
124 practices and facilitate checks on operating activities by regulators (Cerutti et al., 2008; Ezzine de Blas
125 and Ruiz Pérez, 2008; Karsenty et al., 2008; Putz et al., 2008a).

126 For local development, FMPs require that concessions adhere to “social contracts”, redistributing part
127 of the benefits to the local population, either through specific forest taxation or the direct funding of
128 local infrastructure (for example, companies often build wood-processing facilities, such as sawmills,
129 that employ local workers; ATIBT, 2007).

130 In all of the Congo Basin countries except the CAR,¹ the FMP is established by the logging company on
131 the basis of national standards and under the control of forest administrations. After the attribution
132 of forest concessions, logging companies can start logging immediately but have to prepare their FMP
133 within a maximum of three years. The FMP is then reviewed by the forest administration, which
134 evaluates the quality of the plan and either approves it or sends it back to the company with a request
135 for review. In practice, this three-year period is poorly-respected. Moreover, FMPs may not deliver the
136 expected outcomes. First, logging concessions are responsible for the drafting of the FMP, which will
137 thus best fit their strategy: the FMP proposed by the owner of the logging concession will reflect the
138 relative weight they put on conservation and economic outcomes (Cerutti et al., 2017). Second, the
139 fact that an officially-approved FMP exists is neither a quality guarantee nor an indication of its
140 implementation on the ground (Karsenty et al., 2017).

141 **2.2 FSC certification: an additional guarantee of sustainable forest management**

142 To show their commitment toward sustainable forest management, logging companies with an
143 accepted FMP can apply to be certified by the Forest Stewardship Council (FSC). This is a voluntary,
144 market-based approach to enhancing sustainable forest management. Concessions with FSC

¹ CAR is the only country in the Congo Basin where a public structure carries out the FMP for logging companies, mainly because the CAR has since 2000 benefited from a support project for the implementation of FMPs (the PARPAF project financed by the AFD).

145 certification commit to comply to FSC standards, which aim to promote “*environmentally appropriate,*
146 *socially beneficial and economically viable management of the world’s forests*” (FSC, 2018). In return,
147 the FSC label on the forest’s products is expected to be beneficial in terms of market access and share,
148 and higher prices (Romero et al., 2017). For certification, concessions commit to adhere to the ten
149 international FSC principles and twelve criteria, covering social aspects such as workers' rights and
150 employment conditions, and environmental aspects, including diverse measures of forest-
151 management planning and monitoring similar to those that are supposed to appear in their FMP.
152 Independent certifying bodies audit concessions prior to certification to determine their conformity to
153 the FSC criteria: they then provide certification for five years, during which they carry out annual
154 concession inspections to ensure their continued compliance (FSC, 2018).

155 In the context of weak developing-country institutions in, where regulators have limited resources to
156 enforce compliance to Forestry Law and FMP, this third-party verification should provide additional
157 guarantees that logging concessions have effectively adopted sustainable forest-management
158 practices (Blackman et al., 2018). For this reason, regarding the environmental aspects of forest
159 management, the added value of the FSC is to avoid FMPs that only reflect economic criteria and apply
160 only on paper, with few, or no, measures implemented in practice.

161 **2.3 Theory of change**

162 >>>> HERE Figure 1. Theory of change <<<<

163 Figure 1 summarizes the theory of change through which the adoption of sustainable forest-
164 management practices via FMP and FSC is supposed to reduce deforestation in logging concessions.
165 FMP and FSC can have a variety of impacts, including social and economic benefits and reduced forest
166 degradation, which are likely correlated with deforestation. However, the exact measurement of them
167 is beyond the scope of our work here, which will focus only on deforestation. Our theoretical
168 framework is then articulated around five main causal pathways relating forest management to
169 deforestation: (i) concession planning; (ii) monitoring of the concession for settlement expansion,

170 agriculture expansion and illegal activities; (iii) planning of the logging-track network, log landings and
171 skid trails; (iv) improvements in forestry-management practices and logging techniques; and (v)
172 improved livelihoods for local communities (Cerutti et al., 2017; Durrieu de Madron et al., 2011; Ezzine
173 de Blas and Ruiz Pérez, 2008; Pearson et al., 2014; Putz et al., 2008a, 2008b).

174 The FMP first allows logging firms to plan their activity over time, by dividing the concession into
175 different management series and through the production of forest inventories. Moreover,
176 participatory mapping activities with local communities help identify the areas of the concession
177 devoted to community development and small-scale agriculture. These activities could help reduce
178 deforestation in different ways. In production series, rotation planning and the definition of annual cut
179 areas should reduce the expansion, dispersion and sprawl of logging activities, while ensuring that the
180 forest remains undisturbed between exploitation cycles, thereby reducing the repeated exploitation
181 of the same areas. In addition, the definition of conservation series and buffer zones in more
182 vulnerable areas should increase the area that is not logged and thus is without new logging roads.
183 Last, the definition of community-development series should limit forest clearing for agricultural
184 activities and settlement expansion in predefined areas.

185 Second, FMPs involve concession monitoring in order to control the expansion of settlements and
186 agricultural areas, as well as illegal activities. This includes controlling concession access: the
187 temporary or permanent closure of logging tracks, the dismantling of bridges and post-exploitation
188 access control. This monitoring is expected to reduce illegal activities such as slash and burn
189 agriculture, hunting and the illegal harvesting of timber or fuelwood, which could produce
190 deforestation through forest clearing, repeated forest exploitation or even fire spread.

191 Third, FMPs involve the planning of logging tracks, log landings and skid trails. The main technical
192 intervention here is the planning and optimization of the track network according to the topography,
193 forest inventories and the location of annual cut areas in order to preserve soil and valuable forest
194 species for biodiversity and future exploitation. The objectives are to reduce the areas occupied by

195 logging tracks, log landings and skid trails. This is expected to reduce deforestation and the damage to
196 forest cover linked to logging.

197 Fourth, FMPs involve the adoption of a set of improved forestry-management practices and logging
198 techniques, mainly (i) the application of a minimum log diameter (over the legal minimum) that should
199 reduce the volume and increase the variety of logged species, reducing the pressure on the individual
200 most-valuable species and (ii) the improvement of tree-felling techniques (controlled or directional
201 tree felling) which should limit the damage to the remaining stand linked to tree fall and skidding
202 manoeuvres. These practices are mostly expected to affect forest degradation, but should also reduce
203 deforestation by preventing large canopy gaps and tree-felling in sensitive areas that may require long
204 recovery times.

205 Finally, through the associated social measures, FMPs could enhance the livelihoods of those who live
206 and work in and around logging concessions. Improved livelihoods in turn may reduce the incentives
207 for both illegal and unsustainable logging, and could also reduce clearings by reducing the dependence
208 on fuelwood and slash and burn agriculture. However, the relationship between livelihoods and
209 deforestation is complex and, in some cases, improved livelihoods may spur forest-cover change or
210 attract more people (Chomitz, 2007; Rist et al., 2011), potentially increasing deforestation (Blackman
211 et al., 2018).

212 FSC certification is hypothesized to affect deforestation through the same causal mechanisms as noted
213 above. In addition, FSC certification should also enhance monitoring by external actors, including
214 independent certifying bodies, NGOs and the media (Blackman et al., 2018). In the context of weak
215 governance, this should result in better compliance with and performance of each of these
216 mechanisms. To the extent that the enforcement of sustainable forest-management practices by
217 regulators in the study area is weak, we may expect to find a greater fall in deforestation in concessions
218 that are FSC-certified.

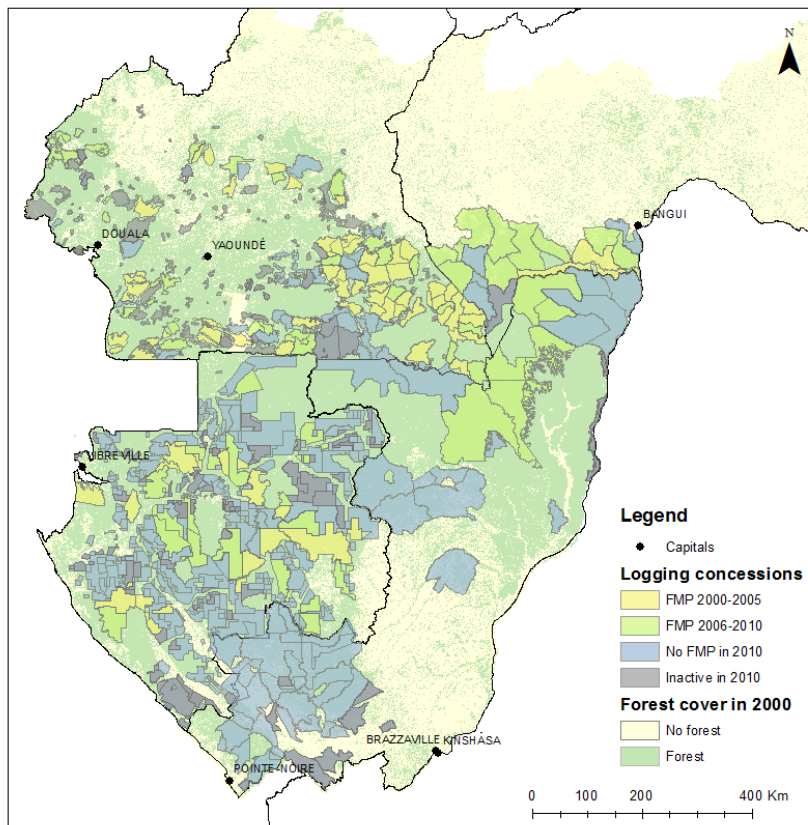
219 By their nature, these mechanisms are likely to produce effects over different time frames and in
220 distinct areas inside concessions. At first, the planning and monitoring of concessions, as well as
221 improved livelihoods, would likely produce effects that are visible in the short to medium term in areas
222 close to settlements, the main transport networks and previously-opened logging roads. In the same
223 timeframe, the planning of logging tracks and log landings is expected to affect the forest in production
224 series through the enforcement of annual cut areas. In the second, more distant, period the adoption
225 of improved forestry-management practices and logging techniques is also expected to affect the
226 forest in production series by allowing valuable trees to regenerate. For these reasons, the impact of
227 sustainable forest-management practices on deforestation should vary over both time and space
228 within concessions with FMPs or FSC certificates.

229 3 Datasets

230 We here use two types of information to evaluate the effect of sustainable forest-management
231 practices promoted via FMP and FSC.

232 We initially collected detailed information on logging concessions in the study area using the official
233 land-tenure data released by the OFAC and World Resources Institute (WRI) in the "Congo Basin Forest
234 Atlases". The datasets used in this study cover 397 concessions across the four countries under
235 consideration (see Map 1). The resulting database was updated using the gray literature and
236 information collected on the ground from local actors, especially in the case of concession reallocation
237 to another company during the study period.

238 Map 1. Location of concessions in the countries analysed in the Congo Basin



239

240 To establish when a logging concession started implementing its FMP, we rely on the FMP-acceptance
 241 date, despite there being potentially long delays between FMP preparation, submission and
 242 acceptance by the competent authorities. We likewise used the issuance date of the FSC certificate to
 243 identify logging concessions whose practices have been verified and certified by an FSC-accredited
 244 external agent. As logging concessions may introduce sustainable forest-management practices ahead
 245 of FMP validation, we will underestimate the FMP effect as some no-FMP concessions in 2010 will
 246 already have a FMP in action. We explore some of these implications in Section 6 when considering
 247 the limits of our work. Other information collected on logging concessions include the physical
 248 attributes of their environment (altitude, steepness and biomass) and their proximity to road
 249 infrastructures and settlements, which can affect both the likelihood of adopting sustainable forest-
 250 management practices and competition over forest resources and deforestation.

251 The second type of information comes from high-resolution maps of forest cover and forest-cover
 252 changes across the Congo Basin. These come from two sources. We first requested and obtained the

253 original maps produced as part of the global effort to reduce emissions from deforestation and forest
 254 degradation in the Congo Basin. To quantitatively assess the spatial and temporal dynamics of forest
 255 change, the governments of Cameroon, CAR, Congo and Gabon developed national forest-monitoring
 256 systems (NFMS). As part of this programme, a number of remote-sensing projects were carried out in
 257 each of these countries in close collaboration with the administration in charge of forest monitoring.
 258 The resulting maps are based on high-resolution satellite imagery and ground-verification data, and
 259 should provide greater cartographic and thematic accuracy than global data (Sannier et al., 2016).
 260 Combining these data, we produced homogeneous regional-level maps of forest cover at three points
 261 in time (1990, 2000 and 2010) and calculated gross deforestation between these dates (See Table 1
 262 and the map in the SI).

263 Second, for comparison purposes, we use measures of tree-cover loss produced from the Global Forest
 264 Change (GFC) dataset (1.0) (Hansen et al., 2013). We calculated tree-cover loss between 2000 and 2010
 265 for two tree-cover thresholds, 30% and 70%. The 30% tree-cover threshold is that used in most forest
 266 definitions, but in the case of the countries of the Congo Basin, the 70% tree-cover threshold seems to
 267 be more realistic given the forest conditions on the ground (Sannier et al., 2016).

268 Table 1. Forest cover and forest-cover change in the study area

Country	Period	Forest cover (km ²)	Deforested area (km ²)	Deforestation rate (%)
Congo	1990-2000	223 554	1 375	0.62
	2000-2010	233 595	1 911	0.82
Gabon	1990-2000	237 242	1 025	0.43
	2000-2010	236 634	512	0.22
Cameroon	1990-2000	245 396	4 790	1.95
	2000-2010	241 487	4 245	1.76
RCA	1990-2000	98 759	3 140	3.18
	2000-2010	96 364	2 632	2.73
Total	1990-2000	804 951	10 330	1.28
	2000-2010	808 080	9 300	1.15

269 Combining the map giving the location and geographical coverage of each logging concession and its
 270 sustainable forest-management practice status to the high-resolution deforestation maps informs us
 271 about the deforested area over 1990-2000 and 2000-2010 in each concession. However, the raw
 272 comparison of the area deforested to time of FMP-acceptance or FSC certificate-issuance is
 273 unsatisfactory for at least two reasons. First, logging concessions had their FMP accepted and received

274 their FSC certificates at different points in time. Hence, in line with the theory of change, we need to
275 take the appropriate definition according to the treatments in which we are interested. Second, the
276 decision to adopt sustainable forest-management practices and submit an FMP is initiated by the
277 logging companies, and is thus to some extent endogenous. The concessions that chose to adopt
278 sustainable forest-management practices likely differ from those that did not, and these differences
279 can affect deforestation. There is thus selection bias in the raw comparisons of logging concessions
280 with and without an FMP, so that we risk attributing the effect of other observable or unobservable
281 concession characteristics to sustainable forest-management practices. The next section describes the
282 empirical framework used to address this problem and select concessions based on the likelihood that
283 the effects of their activities contribute to the deforestation measured over the observation periods.
284 We then present the potential-outcomes framework of Rubin (1974) that we use to deal with potential
285 confounders and estimate the deforestation effect of sustainable forest-management practices.

286 4 Empirical framework

287 Following the theory of change outlined above, we wish to evaluate how deforestation in a concession
288 changes with the adoption of sustainable forest-management practices, measured either by FMP-
289 acceptance or FSC certificate-issuance. We would furthermore like to differentiate the short- and
290 medium- to long-term impacts of sustainable forest-management practices. Finally, we will look for
291 spatial heterogeneity in the average treatment effects.

292 4.1 Treatment groups

293 The first logging concession in the study area had its FMP accepted in 1999. We hence focus on the
294 impact of (i) having an FMP accepted between 2000 and 2005, (ii) having an FMP accepted between
295 2006 and 2010 and (iii) obtaining an FSC certificate between 2000 and 2010 on deforestation between
296 2000 and 2010.

297 Measuring the effect of the early adoption of sustainable forest-management practices (treatment
298 *FMP 2000-2005*) reflects the potential FMP impact on deforestation over the medium to long run. We
299 expect the concessions with an accepted FMP before 2005 to adopt selective logging practices over at
300 least five years, so that deforestation between 2000 and 2010 will be lower than in concessions
301 without an FMP over this period (See Figure S1). However, as very few concessions had an accepted
302 FMP in 1999, our data do not allow us to measure the impact of FMPs over longer time periods.

303 We next consider more treated concessions, defined as those that had an FMP accepted between 2006
304 and 2010 (treatment *FMP 2006-2010*). As deforestation is measured in 2010, this treatment reflects
305 the short term, and supposes that logging companies began improving their forest management
306 before FMP acceptance, as otherwise the time period is too short for us to observe a reduction in
307 deforestation. There may be a long delay between FMP preparation, submission and acceptance by
308 the competent authorities, and concessions may begin to implement FMP activities before its
309 acceptance. The effects of the FMP may thus already be apparent in 2010.

310 In both of these two treatments, the control group is active concessions without an FMP. We define a
311 concession as "active" if it was attributed to a logging company for at least two years for the *FMP 2000-*
312 *2010* treatment (i.e. since 2008) and at least five years for the *FMP 2000-2005* treatment (i.e. since
313 2005, in order to be consistent with the treated concessions that, by definition, have all been active
314 since 2005). The "no-FMP concessions" hence include all the active concessions that had no FMP in
315 2010 (in 2005, respectively, for the *FMP 2000-2005* treatment), including concessions with accepted
316 FMP after 2010 or that had an FMP in process in 2010. For the *FMP 2000-2005* treatment, concessions
317 that had an FMP accepted between 2005 and 2010 were excluded.

318 Overall, there are 60 FMP concessions and 166 no-FMP concessions for the *FMP 2000-2005* treatment
319 and 121 FMP concessions and 194 no-FMP concessions for the *FMP 2000-2010* treatment (see
320 Appendix 1 for more details).

321 Despite the certification of sustainable forest-management practices being recent in the Congo Basin,
322 with the first certificates only issued in 2005, we can estimate the impact of FSC certification (the *FSC*
323 *2000-2010* treatment) on 2000-2010 deforestation. Since the first FSC certificates were issued in 2005,
324 we here evaluate the short-term impact of FSC certification (after one to five years of certification). It
325 is however worth noting that all FSC-certified concessions already had a valid FMP. Furthermore, over
326 half of the concessions with FSC certificates had an accepted FMP before 2005. As such, estimating the
327 effect of FSC-certificate issuance is similar to measuring the impact of an FMP, but with these particular
328 logging concessions in addition benefiting from third-party verification of sustainable forest-
329 management practices. The treated group here is all active concessions that were certified before
330 2010. As in the previous treatments, the control group is all active concessions without an FMP in 2010.
331 There are 25 FSC concessions and 194 no-FMP concessions in this treatment.

332 **4.2 Econometrics and identification strategy**

333 This subsection describes the strategy used to account for the endogenous selection of logging
334 concessions into the adoption of sustainable forest-management practices described in Section 3. Our
335 approach here is consistent with the previous empirical literature on the environmental impact of
336 various policies (see for instance Blackman, 2013; Börner et al., 2016; Le Velly and Dutilly, 2016) and
337 uses a propensity-score matching (PSM) approach to estimate the effect of FMP and FSC-certification
338 in the Congo Basin with the least possible bias.

339 Using the potential-outcome framework, we consider that each logging concession has two potential
340 outcomes Y_1 and Y_0 , where Y_1 is the area deforested between 2000 and 2010 for logging concessions
341 with an FMP (or with FSC certification) and Y_0 the analogous figure for concessions without an FMP
342 (FSC certification). T is a dummy for the concession having either an FMP or FSC certification. We want
343 to estimate the average effect of an FMP or FSC certification in the concessions that have them, i.e.
344 the average treatment effect on the treated (ATT):

$$345 \text{ATT} = \tau = E(Y_1 - Y_0 | T = 1)$$

346 As Y_0 is never observed for a “treated” concession, the ATT cannot be directly estimated. Denote by X
 347 a set of characteristics that are known to affect both deforestation and the presence of an accepted
 348 FMP or FSC certificate (which we refer to as the treatment for brevity below). The propensity score is
 349 $p(X) = \Pr(T = 1|X)$ and $p^* = \Pr(T = 1)$ is the probability that a concession be treated. The
 350 following assumptions, often referred to as “strong ignorability” (Rosenbaum and Rubin, 1983), imply
 351 that controlling for X suffices to eliminate all of the confounding factors:

352
$$(H1) Y_1, Y_0 \perp\!\!\!\perp T | X \text{ and } (H2) 0 < p(X) < 1.$$

353 H1 is often referred to as “unconfoundedness”, and states that, if all confounders are included in X ,
 354 then controlling for X renders treatment exposure independent of the potential outcomes. Under H1,
 355 Rosenbaum and Rubin (1983) show that $Y_1, Y_0 \perp\!\!\!\perp T | p(X)$. Consequently, logging concessions with
 356 similar propensity scores would have on average similar deforestation in the absence of an FMP or FSC
 357 Certification and

358
$$E(Y_0|T = 1, p(X)) = E(Y_0|T = 0, p(X))$$

359 H2 implies that, for almost all values of X , both treated and untreated concessions have a probability
 360 of an accepted FMP or FSC certification at some point. If H1 and H2 hold, then Abadie and Imbens
 361 (2016) suggest estimating the ATT τ as follows:

362
$$\hat{\tau} = \frac{1}{N_1} \sum_{i=1}^N T_i \left(Y_i - \frac{1}{M} \sum_{j \in \mathcal{J}_M(i)} Y_j \right)$$

363 Here M is a fixed number of matches per logging concession i , $\mathcal{J}_M(i)$ the set of matches for logging
 364 concession i , N the number of treated and untreated concessions, N_1 the number of concessions with
 365 the treatment and T_i a dummy for the concession i being treated. The matching set $\mathcal{J}_M(i)$ is defined
 366 as follows:

367
$$\mathcal{J}_M(i) = \left\{ j = 1, \dots, N : T_j = 1 - T_i, \left(\sum_{k: T_k = 1 - T_i} \mathbb{1}(|p(X_i) - p(X_k)| \leq |p(X_i) - p(X_j)|) \right) \leq M \right\}$$

368 where $\mathbb{1}(\cdot)$ is an indicator variable for the event inside the brackets holding. The set $J_M(i)$ hence
369 consists of the logging concessions that are not treated and with a propensity score similar to that of
370 logging concession i . Overall, $\hat{\tau}$ is the average difference in the area deforested between each treated
371 concession and the average deforestation in a set of untreated concessions with similar propensity
372 scores. Abadie and Imbens (2016) also show that $\hat{\tau}$ produces an unbiased estimate of the ATT, while
373 taking into account the fact that the propensity score is estimated.

374 **4.3 Confounding factors and estimation**

375 We consider ten key covariates that are known to be correlated with the likelihood of deforestation
376 and the adoption of sustainable forest management to estimate the propensity scores (Blackman,
377 2013). The selected covariates include indicators of accessibility, population pressure, biomass
378 productivity, average steepness and elevation. Four variables were used to proxy various dimensions
379 of accessibility that are the most correlated with deforestation and the likelihood of adopting
380 sustainable forest-management practices: the distance to the road network, the distance to the
381 nearest settlement, distance to the capital of the country and main ports, and the travel distance to a
382 market. Settlement density is the number of settlements in a 20-kilometre radius around each
383 settlement, and picks up population pressure. We also include the distance to a deforested area in the
384 1990-2000 period. Above-ground forest biomass is based on Avitabile et al. (2016) and measures the
385 density of timber available (for example, forests from Southern Congo have less biomass than those in
386 the Northern Congo, where most of the FMP concessions are located). Elevation and slope describe
387 the topographic environment and so suitability for logging, as steep slopes can pose problems for
388 logging machines. Last, we control for the concession area in hectares (see the supplementary
389 information for more information on the covariates).

390 **4.4 Robustness checks**

391 To produce unbiased estimates of the treatment effects, quasi-experimental approaches based on
392 matching techniques assume that all of the relevant variables that can affect both the likelihood of

393 deforestation and the adoption of sustainable forest-management practices are observed and used as
394 controls. However, this assumption is hard to test, as the real unknown variables are by definition
395 unknown, while some known confounders (the quality of local governance) are hard to measure
396 (Panlasigui et al., 2018). If these unobservable confounders are spatially time-invariant, their effect
397 should be seen in the difference in the area deforested in concessions with and without an FMP prior
398 to FMP adoption, and hence between 1990 and 2000. Following this argument, we test for differences
399 in 1990-2000 deforestation between concessions with and without FMP after matching. We
400 furthermore consider an alternative identification approach that explicitly takes into account past
401 deforestation by measuring the effect of FMP adoption on the change in deforestation over time. This
402 change in deforestation (between 1990-2000 and 2000-2010) should in theory allow us to abstract
403 from the effect of any unobservable factors that do not vary over time and hence should not affect the
404 change in deforestation. This is akin to combining matching with a difference-in-difference approach.
405 This is however not our preferred strategy, given that we do not have a true panel of logging
406 concessions. Some logging concessions observed in 2000-2010 were not active in 1990-2000.
407 Moreover, the deforestation data are of poorer quality between 1990 and 2000 due to the lack of
408 satellite imagery, and the GFC dataset only covers deforestation after 2000.

409 **4.5 Impact heterogeneity**

410 To explore the mechanisms of change, we assess impact heterogeneity via pixel-level analyses, which
411 allows us to consider spatial heterogeneity in the average treatment effect inside concessions (see the
412 SI for detailed information on the pixel-sampling strategy). This pixel-level data comes from the high-
413 resolution satellite imagery described in Section 3.

414 To test the most-plausible pathways of the theory of change outlined above, we explore heterogeneity
415 by the proximity of pixels to past deforestation, road networks and settlements (see Table 2 for a
416 summary of the main predictions of the different plausible mechanisms). More precisely, we compare
417 how the difference in deforestation across pixels that are close (under median distance) and far (over

418 median distance) differs by concession FMP status. In line with the theoretical framework, we focus
 419 the heterogeneity analysis on concessions that had their FMP accepted between 2000 and 2005,
 420 where the expected impact of each mechanism is more likely to be seen.

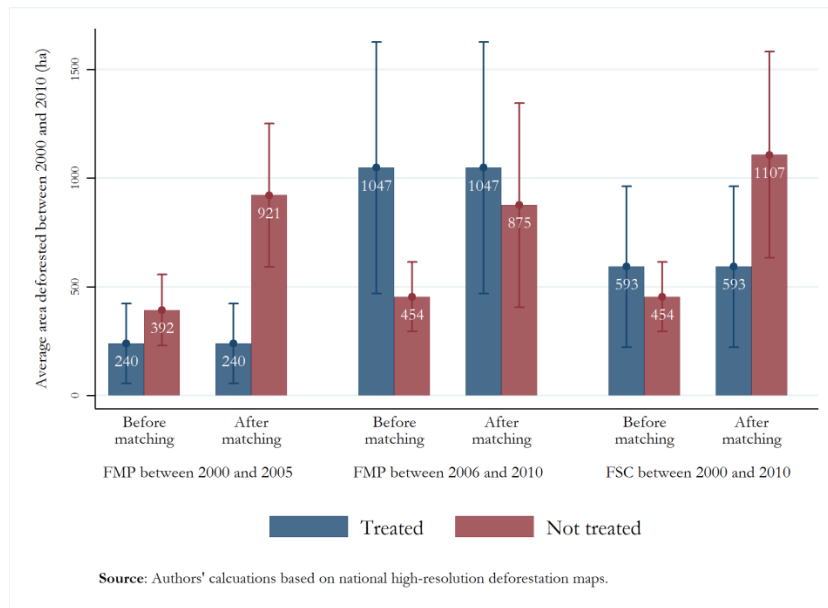
421 Table 2: Predictions of the main falsifiable pathways through which sustainable forest-management
 422 practices can affect deforestation in the short to medium run

Variables tested in the heterogeneity analysis	Mechanism tested	Expected impact
Distance to previous deforestation	Effectiveness of concession planning, especially the mapping of production series	Less deforestation close to previous deforestation due to rotation planning, avoiding the re-exploitation of the areas previously logged
	Effectiveness of concession monitoring, especially control of access by closing former logging roads	Less deforestation close to previous deforestation (due to the opening of logging roads) linked to the reduction of illegal activity along former logging roads
Distance to main transport networks	Effectiveness of concession monitoring with control of access	Less deforestation close to main transport networks due to reduced access from public roads
Distance to settlements	Effectiveness of concession planning, especially the definition of areas for community and agriculture development with the promotion of sustainable activities	Less deforestation close to settlements due to the promotion of sustainable activities and better monitoring of settlement extension
	Effectiveness of concessions' "social contracts"	

423 **5 Results**

424 **5.1 The impact of sustainable forest-management practices on deforestation**

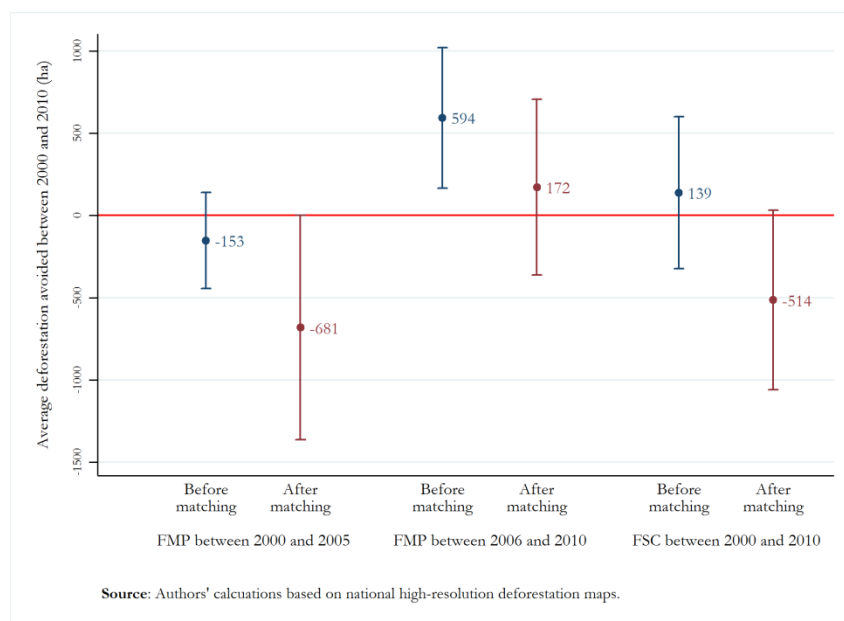
425 Figure 2. Difference before and after matching across treatment groups



426

427

Figure 3. The impact of treatment on 2000-2010 deforestation



428

429 After matching, our estimates suggest that concessions with an accepted FMP between 2000 and 2005
 430 have less deforestation compared to otherwise-similar concessions without an FMP. More precisely,
 431 FMP adoption between 2000 and 2005 is associated with average lower deforestation of 681ha per
 432 concession (Fig. 3). Since the area deforested between 2000 and 2010 is estimated at 921ha in control
 433 concessions, this represents a 74% fall in deforestation (Fig. 2). We find similar results using estimates
 434 of the area deforested from the GFC dataset, with FMP adoption between 2000 and 2005 being

435 associated with lower deforestation of 1005ha for tree cover of 70% and 1144ha for tree cover of 30%,
436 representing respectively drops of 74 and 75% (see the SI).

437 For an accepted FMP between 2006 and 2010, after matching, we find no statistically-significant
438 impact of the FMP 2006-2010 treatment on 2000-2010 deforestation. The same result applies when
439 the area deforested is estimated using tree-cover loss from the GFC dataset for tree cover of 70% and
440 30%. As such, reduced deforestation is not seen in the short run, in line with the predictions from the
441 theory of change.

442 Last, after matching, the FSC 2000-2010 treatment is also associated with a statistically-significant fall
443 in deforestation between 2000 and 2010. Concessions with FSC certification, testifying that sustainable
444 forest-management practices have indeed been implemented, have on average 514ha less
445 deforestation between 2000 and 2010. Compared to the average deforested area of 1107ha in the
446 control concessions, this represents a drop of 48% (Figs. 2 and 3). This result can be replicated using
447 deforestation from the GFC data, with avoided deforestation in FSC 2000-2010 concessions of 699ha
448 for tree cover of 70% (a 47% fall) and 789ha for tree cover of 30% (a 50% fall).

449 **5.2 Robustness checks**

450 The validity of the above results rests on the assumption that the matching was successful in comparing
451 treated and untreated concessions with similar propensity scores. We moreover assume that no
452 variables other than the 10 covariates used as controls predict FMP acceptance and/or FSC certification
453 and deforestation in logging concession. In this subsection we discuss the sensitivity of our estimates
454 to these two assumptions.

455 The matching was first successful in balancing treated and untreated households with similar
456 propensity scores. The Figure in the SI shows that the distribution of propensity scores after matching
457 is identical for treated and the untreated control concessions.

458 However, even after matching, control concessions cover larger tracts of land. The fact that
459 concessions without an FMP cover larger areas than those with an FMP after matching may suggest
460 that our estimate over-estimates the drop in deforestation from the FMP as larger concessions are
461 more likely to have larger areas deforested, even with lower deforestation rates. However, this is not
462 the case: further analyses show that the 2000-2010 deforestation rate is also lower in concessions with
463 an accepted FMP between 2000 and 2005.

464 There is no statistically-significant difference in past deforestation (1990-2000) for concessions with
465 and without an FMP (although concessions with an accepted FMP between 2000 and 2005 exhibited
466 qualitatively less 1990-2000 deforestation).

467 We introduce an alternative specification to account more directly for this 1990-2000 deforestation
468 difference, which may reveal subtle but real differences in unobservable characteristics. This seeks to
469 measure the effect of FMP adoption on the ability of logging concessions to reduce deforestation over
470 time. Comparing the change in deforestation between 1990-2000 and 2000-2010 across logging
471 concessions with and without an FMP, we find that deforestation fell more in treated concessions than
472 in control concessions without an FMP, although this difference was not statistically significant for the
473 FMP 2000-2005 treatment (see the SI). We applied the same approach for our other treatment
474 variables, and found similar statistically-significant results.

475 **5.3 Impact heterogeneity**

476 We first reproduce our analysis at the pixel rather than the previous concession level, and find that
477 pixels located in treated concessions are less likely to be deforested than those in concessions without
478 an FMP, as in the previous Sections.

479 Second, spatial-heterogeneity analysis using the pixel-level database revealed that 2000-2005 FMP
480 produced significantly less deforestation in areas close to settlements and previously-deforested
481 areas, with the measured effect being stronger for observations below the median value of these two
482 variables (see Table 3). The ATT for all concessions was a fall of 0.29 percentage points, equivalent to

483 55% less deforestation; the analogous figures in areas close to settlements are 0.49 (62%) and in areas
 484 close to previous deforestation 0.51 (66%). Conversely, the effect of FMP in areas further from
 485 settlements and areas of previous deforestation was smaller and insignificant. However, we find no
 486 heterogeneity by distance to the main transport networks: the effect is a fall of 0.27 percentage points
 487 close to the transport network and 0.19 percentage points further away, both statistically significant
 488 but similar to the overall ATT figure of 0.29.

489 These results are in line with our expectations from our theory of change (Table 2 and Figure 1). They
 490 emphasize the effects of improvements in, first, the planning of the concessions, especially for rotation
 491 cycles and areas for community and agricultural development, and, second, the monitoring of
 492 concessions by closing former logging roads and monitoring the extension of settlements and
 493 agriculture areas. Only the monitoring of the incursion of public roads into concessions does not
 494 appear to have an effect on the reduction in deforestation.

495 Table 3. The effect of 2000-2005 FMPs on 2010 deforestation: Geographic heterogeneity

	Coef.	Robust std error	P> z	[95% Conf. Interval]	
Overall impact in concessions	-0.00285***	0.00057	0.0000	-0.00398	-0.00173
Heterogeneity:					
Closer to settlements (< 10km)	-0.00486***	0.00098	0.0000	-0.00677	-0.00677
Further from settlements (>= 10km)	-0.00059	0.00070	0.3990	-0.00197	0.00078
Closer to previous deforestation (< 4km)	-0.00508***	0.00088	0.0000	-0.00681	-0.00336
Further from previous deforestation (>= 4km)	0.00004	0.00064	0.9480	-0.00120	0.00129
Closer to main transport network (< 15km)	-0.00268***	0.00103	0.0090	-0.00470	-0.00066
Further from main transport network (>= 15km)	-0.00189***	0.00068	0.0050	-0.00321	-0.00056

496 *Note:* *p<0.1, **p<0.05, ***p<0.01.

497 6 Discussion and concluding remarks

498 Curbing tropical deforestation is arguably one of the main environmental challenges. Addressing it
499 requires the assessment of policy effectiveness and the understanding of the mechanisms
500 underpinning their successes and failures. This paper contributes to this aim by showing that the area
501 deforested is lower in logging concessions that adopt sustainable forest-management practices in the
502 Congo Basin. Deforestation is lower in concessions that have had an FMP for at least five years. Like
503 Panlasigui et al. (2018), this highlights the importance of the time frame: interventions aimed at
504 increasing FMPs and FSC-adoption should be evaluated over long time periods.

505 Evidence from micro-level analyses suggests that FMP have allowed concessions to avoid the
506 overexploitation of previously-logged areas. Our results also suggest that FMP concessions are more
507 likely to better control access into their perimeter and reduce deforestation around communities
508 located within or nearby the concession. This is in line with the theory of change underpinning the
509 adoption of sustainable forest-management practices. These results confirm the utility of potential
510 spatial heterogeneity in policy and management interventions (Bruggeman et al., 2018).

511 While FMP acceptance is mandatory across countries in the Congo Basin, logging concessions chose
512 when to draft and submit their FMP. It is then possible that concessions that had their FMP accepted
513 earlier have unobserved characteristics that led them also to deforest less. Our efforts to account for
514 this were limited by the fact that logging concessions change ownership over time, and that
515 information about the former management was scarce. However, taking into account previous
516 deforestation, we found that the area deforested fell more in concessions following the FMP adoption.
517 Whether deforestation will also be lower in logging concessions that had their FMP accepted later
518 remains an open question. Will we continue to see lower 2005-2015 deforestation in concessions with
519 an FMP accepted between 2005 and 2010? Will there continue to be lower deforestation in
520 concessions that had their FMP accepted earlier?

521 Answering the above questions is a natural extension of our work here, and will help address the
522 external validity of our results. This will also help inform us whether the adoption of sustainable forest-
523 management practices works for all concessions, and how lower deforestation varies over longer time
524 periods. Likewise, the adoption of sustainable forest-management practices is also expected to bring
525 benefits other than reduced deforestation. These include, for example, conservation values such as
526 reducing forest degradation, the preservation of biodiversity, and welfare improvements for the local
527 population. Future work should therefore address other potential FMP impacts in the Congo Basin,
528 and reveal whether lower deforestation has come at the expense of other dimensions of development
529 and conservation

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Figure 1. Theory of change

